

TOXICOLOGICAL REVIEW

OF

HYDROGEN SULFIDE

(CAS No. 7783-06-4)

In Support of Summary Information on the Integrated Risk Information System (IRIS)

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FOREWORD

The purpose of this Toxicological Review is to provide scientific support and rationale for the hazard and dose-response assessment in IRIS pertaining to chronic exposure to hydrogen sulfide. It is not intended to be a comprehensive treatise on the chemical or toxicological nature of hydrogen sulfide.

In Section 6, Environmental Protection Agency (EPA) has characterized its overall confidence in the quantitative and qualitative aspects of hazard and dose response. Matters considered in this characterization include knowledge gaps, uncertainties, quality of data, and scientific controversies. This characterization is presented in an effort to make apparent the limitations of the assessment and to aid and guide the risk assessor in the ensuing steps of the risk assessment process.

For other general information about this assessment or other questions relating to IRIS, the reader is referred to EPA's IRIS Hotline at 202-566-1676.

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This document and summary information on IRIS have received peer review both by EPA scientists and by independent scientists external to EPA. Subsequent to external review and incorporation of comments, this assessment has undergone an Agency-wide review process whereby the IRIS Program Director has achieved a consensus approval among the Office of Research and Development; Office of Air and Radiation; Office of Prevention, Pesticides, and Toxic Substances; Office of Solid Waste and Emergency Response; Office of Water; Office of Policy, Economics, and Innovation; Office of Children's Health Protection; Office of Environmental Information; and the Regional Offices.

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Summaries of the external peer reviewers' comments and the disposition of their recommendations are in Appendix A.

1. INTRODUCTION

This document presents background and justification for the hazard and dose-response assessment summaries in EPA's Integrated Risk Information System (IRIS). IRIS summaries may include an oral reference dose (RfD), inhalation reference concentration (RfC), and a carcinogenicity assessment.

The RfD and RfC provide quantitative information for noncancer dose-response assessments. The RfD is based on the assumption that thresholds exist for certain toxic effects such as cellular necrosis but may not exist for other toxic effects such as some carcinogenic responses. It is expressed in units of mg/kg-day. In general, the RfD is an estimate (with uncertainty spanning perhaps an order of magnitude) of a daily exposure to the human population (including sensitive subgroups) that is likely to be without an appreciable risk of deleterious, noncancer effects during a lifetime. The inhalation RfC is analogous to the oral RfD, but provides a continuous inhalation exposure estimate. The inhalation RfC considers toxic effects for both the respiratory system (portal-of-entry) and for effects peripheral to the respiratory system (extrarespiratory or systemic effects). It is generally expressed in units of mg/m³.

The carcinogenicity assessment provides information on the carcinogenic hazard potential of the substance in question and quantitative estimates of risk from oral exposure and inhalation exposure. The information includes a weight-of-evidence judgment of the likelihood that the agent is a human carcinogen and the conditions under which the carcinogenic effects may be expressed. Quantitative risk estimates are presented in three ways. The *slope factor* is the result of application of a low-dose extrapolation procedure and is presented as the risk per mg/kg-day. The *unit risk* is the quantitative estimate in terms of either risk per μ g/L drinking water or risk per μ g/m³ air breathed. Another form in which risk is presented is a drinking water or air concentration providing cancer risks of 1 in 10,000; 1 in 100,000; or 1 in 1,000,000.

Development of these hazard identification and dose-response assessments for hydrogen sulfide has followed the general guidelines for risk assessment as set forth by the National Research Council (1983). EPA guidelines that were used in the development of this assessment may include the following: the *Guidelines for the Health Risk Assessment of Chemical Mixtures* (U.S. EPA, 1986a), *Guidelines for Mutagenicity Risk Assessment* (U.S. EPA, 1986b), *Guidelines for Developmental Toxicity Risk Assessment* (U.S. EPA, 1991), *Draft Revised Guidelines for* Carcinogen Risk Assessment (1999), Guidelines for Reproductive Toxicity Risk Assessment (U.S. EPA, 1996), Guidelines for Neurotoxicity Risk Assessment (U.S. EPA, 1998); Recommend-ations for and Documentation of Biological Values for Use in Risk Assessment (U.S. EPA, 1988); (proposed) Interim Policy for Particle Size and Limit Concentration Issues in Inhalation Toxicity (U.S. EPA, 1994a); Methods for Derivation of Inhalation Reference Concentrations and Application of Inhalation Dosimetry (U.S. EPA, 1994b); Peer Review and Peer Involvement at the U.S. Environmental Protection Agency (U.S. EPA, 1994c); Use of the Benchmark Dose Approach in Health Risk Assessment (U.S. EPA, 1995b); Science Policy Council Handbook: Peer Review (U.S. EPA, 2000a); Memorandum from EPA Administrator, Carol Browner, dated March 21, 1995, Subject: Guidance on Risk Characterization; and Science Policy Council Handbook: Risk Characterization (U.S.EPA, 2000b).

Literature search strategies employed for this compound were based on the CAS Registry Number and at least one common name. At a minimum, the following databases were searched: HSDB, TSCATS, CCRIS, GENETOX, EMIC, EMICBACK, DART, ETICBACK, TOXLINE, CANCERLINE, MEDLINE, and MEDLINE backfiles. The relevant literature was reviewed through 2002. Pertinent scientific information submitted by the public to the IRIS Submission Desk was also considered in the development of this document.

Previous IRIS entries for hydrogen sulfide include an RfD listed as 3×10^{-3} mg/kg-day based on an oral pig study in which the critical endpoint was GI disturbance (Wetterau et al., 1964). This figure is based on a NOAEL of 3.1 mg/kg-day and a LOAEL of 15 mg/kg-day with an uncertainty factor of 1,000. The RfC for H₂S was previously listed as $1 \times 10^{-3} \text{ mg/m}^3$ based on the mouse subchronic inhalation study by CIIT (1983a) using inflammation of the nasal mucosa as the critical endpoint. The derivation of the RfC used the NOAEL_{HEC} of 1 mg/m^3 and the LOAEL_{HEC} of 2.6 mg/m^3 , as well as an uncertainty factor of 1,000. The cancer assessment for hydrogen sulfide had not yet undergone a complete evaluation and determination under US EPA's IRIS program for evidence of human carcinogenic potential.

2. CHEMICAL AND PHYSICAL INFORMATION RELEVANT TO ASSESSMENTS

Hydrogen sulfide is known as dihydrogen monosulfide, dihydrogen sulfide, hydrosulfuric acid, sewer gas, stink damp, sulfureted hydrogen, and sulfur hydride. Some relevant chemical

and physical properties of hydrogen sulfide are listed below, National Library of Medicine, Hazardous Substances Data Bank (HSDB, 1998).

CAS registry number:	7783-06-4
Molecular formula:	H_2S
Molecular weight:	34.08
Vapor pressure:	15,600 mm Hg at 25°C
Density:	1.5392 g/L at 0°C, 760 mm Hg
Boiling point:	-60.33°C
Water solubility:	3980 mg/L at 20°C
Log K _{ow}	Not applicable
Dissociation constants:	pKa1 = 7.04; pKa2 = 11.96
Conversion factor:	$1 \text{ ppm} = 1.39 \text{ mg/m}^3$

Hydrogen sulfide is a colorless gas and has a strong odor of rotten eggs (HSDB, 1998). Hydrogen sulfide is soluble in certain polar organic solvents, notably methanol, acetone, propylene carbonate, sulfolane, tributyl phosphate, various glycols, and glycol ethers (HSDB, 1998). It is also soluble in glycerol, gasoline, kerosene, carbon disulfide, and crude oil. Aqueous solutions of H_2S are not stable; absorbed oxygen causes the formation of elemental sulfur and the solutions become turbid rapidly (HSDB, 1998).

The primary uses of H_2S include the production of elemental sulfur and sulfuric acid, the manufacture of heavy water and other chemicals; in metallurgy; and as an analytical reagent (HSDB, 1998). In agriculture, it is used as a disinfectant. Formulations include a technical grade (98.5%) and a purified grade (99.5% min.). Occupational exposure to H_2S occurs primarily from its presence in petroleum, natural gas, soil, sewer gas and as a byproduct of chemical reactions, e.g., viscose rayon and certain leather tanning processes.

Hydrogen sulfide gas is a natural product of decaying organic matter. In residential settings it is most commonly the result of decomposition in septic or sewer systems.

3. TOXICOKINETICS RELATIVE TO ASSESSMENTS

3.1. ABSORPTION

No studies were identified which quantitatively determined the absorption of H_2S in humans or animals. However, the inhalation of high concentrations of H_2S (greater than 2000 ppm or 2780 mg/m³) can be fatal within seconds or minutes in both humans and animals, suggesting that it is absorbed rapidly through the lungs. Studies in pigs and rats measuring plasma sulfide levels suggested that H_2S is absorbed following ingestion. Since H_2S exists as a

gas, oral exposure is not likely to occur. At physiological pH 7.4 approximately one third of the H_2S exists in the undissociated form and the remainder largely as the hydrosulfide anion (Beauchamp et al., 1984). Although hydrogen is reported elsewhere to be fairly water soluble, Milby and Baselt (1999a) indicated that H_2S is only moderately water soluble and thus not well-absorbed by the moist tissues of the upper respiratory tract. The percutaneous absorption of H_2S has been described as slow (Laug and Draize, 1942).

3.2. METABOLISM

Hydrogen sulfide is metabolized by three pathways: oxidation, methylation, and reactions with metalloproteins. However, the major pathway is oxidation in the liver with methylation being a very minor pathway. The major oxidation product of H_2S metabolism is sulfate (Beauchamp et al., 1984) although thiosulfate may also be isolated from tissues exposed to hydrogen sulfide (Kage et al., 1992). At physiological pH (pH-7.4), H_2S is approximately 69% ionized (HS⁻). The metabolic pathway for H_2S is depicted in Figure 1.

Bartholomew et al. (1980) perfused heparinized blood containing sodium [S³⁵]sulfide through isolated rat lungs, liver, and kidney and determined the rate of oxidation for each organ. In the presence of water, sodium sulfide is rapidly converted to H_2S (Haggard, 1921). Therefore, administration of sodium sulfide in the isolated perfused system would be similar to inhalation exposure to H_2S . In the isolated perfused liver, approximately 70% sulfide was rapidly metabolized to sulfate within 15 minutes, and 82% of the detected radioactivity was sulfate after 2 hours perfusion. When unlabeled thiosulfate was added to the liver perfusion system, 54% of the radioactivity was thiosulfate and 22% was sulfate after 15 minutes of perfusion. After 2 hours, these proportions reversed. This would suggest that sulfide is oxidized to thiosulfate with further oxidation to sulfate.

In the isolated perfused lung, 32% of the administered radioactivity was lost from the blood as volatile ³⁵S after 15 minutes. However, after 15 minutes no further radioactivity was lost. Sulfide was oxidized slowly to thiosulfate, and only trace amounts of sulfate were present after 2 hours of perfusion. The small amount of detectable sulfate in the isolated perfused lung was thought to be due to the absence of sulfide oxidase in the lung. In the kidney, sulfide was slowly oxidized to sulfate possibly through thiosulfate as an intermediate metabolite. The investigators also reported that the rat liver mitochondria catalyzed the oxidation of sulfide to thiosulfate *in vitro* through an unknown mechanism. This oxidation was inhibited at high concentrations of sulfide (60µM) because of reduced oxygen consumption by the mitochondria

due to cytochrome *c* oxidase inhibition. When glutathione (GSH) was added to the rat liver mitochondria *in vitro*, thiosulfate was further metabolized to sulfate.

Kage et al. (1992) exposed adult male Japanese white rabbits to 100-200 ppm (139-278 mg/m³) for 60 minutes and to 500-1000 ppm (695-1390 mg/m³) until fatal (22 minutes, approx.). In the low-level exposure group, blood and urine were collected immediately after exposure and 1, 2, 4, 6, and 24 hours following exposure. Blood and organs were collected immediately after death in high-exposure animals. In the high-exposure animals, thiosulfate levels in the blood, lung, and brain were 2-7 times higher than sulfide levels. Thiosulfate and sulfide were absent or present only in trace amounts in the liver, kidney, and muscle. In the low-exposure animals, blood thiosulfate decreased rapidly to trace levels 2 hours after exposure and was not detected 4 hours after exposure. Urinary thiosulfate levels exhibited a peak at one to two hours followed by a decline, but were still detectable 24 hours after exposure. Sulfide was not detected in the blood or urine of low-exposure animals.

It appears that oxidation is also the major metabolic pathway for H_2S in humans. Thiosulfate was found in the urine of volunteers exposed to 8, 18, or 30 ppm (11, 25, or 42 mg/m³) H_2S for 30-45 minutes (Kangas and Savolainen, 1987). They also determined thiosulfate levels in maintenance workers in a pelt processing plant with known exposures to H_2S , which were compared with thiosulfate levels in unexposed workers. Urinary thiosulfate concentration increased according to the gas concentration and exposure time. The highest concentration of urinary thiosulfate in the exposed workers was observed 15 hours after exposure. At 17 hours, urinary levels of thiosulfate had returned to control levels suggesting that all of the absorbed sulfide had been oxidized within the 15 hours after exposure.

The relationship between sulfide concentration and cytochrome oxidase in target tissues was explored by exposing male CD rats (6/group) to H₂S at 0, 10, 30, 80, 200, or 400 ppm (0, 14, 42, 111, 278, or 556 mg/m³) once for 3 hr and for 70 days (Dorman et al., 2002). The only target organ that demonstrated a clear dose response relationship with increased sulfide levels and concomitant decreases in cytochrome oxidase was the lung examined at the end of the single 3-hour exposure; statistical significance was achieved at 80 ppm (111 mg/m³) and above for increases in sulfide and at 30 ppm (42 mg/m³) and above for decreases in cytochrome oxidase. Sulfide concentrations in the lung decreased to preexposure levels within 15 min of the end of exposure. The response was not seen in the lung after the 70-day repeated exposure. Hindbrain sulfide concentrations and cytochrome oxidase activities were unaffected by single or repeated exposures. Cytochrome oxidase concentrations were decreased ($p \le 0.05$) in both the nasal respiratory and olfactory epithelium at levels of 30 ppm (42 mg/m³) although sulfide levels did

not appear to be decreased until 400 ppm (556 mg/m³) H_2S . These results indicate that cytochrome oxidase is a sensitive biomarker of effect for H_2S exposure (more so than sulfide levels) for both lung and nasal epithelial tissues, but only for short term exposures.

Methylation has also been reported as a possible metabolic pathway for H_2S . Hydrogen sulfide was reported to be methylated to methanethiol *in vitro* by the intestinal mucosa of Sprague-Dawley rats (Weisiger et al., 1980). Methanethiol can be further methylated, although much more slowly, to dimethylsulfide. Thiol *S*-methyltransferase was reported to catalyze the two reactions. It is a ubiquitous enzyme, and the investigators reported the highest activity in the colonic and cecal mucosa (10^{-13} mol/min/mg of protein). High activities were also reported in the liver, lung, and kidney.

Levitt et al. (1999) also studied the metabolism of H_2S in the cecal mucosa of Sprague-Dawley rats and concluded that oxidation, rather than methylation, was the primary metabolic route in the cecal mucosa. When incubated with H_2S , cecal tissues did not produce methanethiol or dimethylsulfide even when a methyl donor, *S*-adenosylmethionine, was added to the reaction mixture. The rates of metabolism for cecal and liver homogenates were calculated to be 21 and 2.5 nmol/min/mg of protein, respectively, and the metabolic products of the cecal and liver homogenate were reported to be sulfate and thiosulfate, the primary metabolite. The oxidation rate is approximately 10,000 times the methylation rate reported by Weisiger et al. (1980), which would suggest that methylation is not an important pathway for H_2S detoxification.

Another possible pathway of H_2S metabolism is conjugation with GSH. Smith and Abbanat (1966) reported that exogenously-oxidized, but not reduced, GSH protected mice from lethal intraperitoneal (i.p.) injections of sodium sulfide. They also stated that endogenous oxidized GSH, and other endogenous compounds containing disulfide bridges might provide an important detoxification process following H_2S exposure.

Hydrogen sulfide has also been reported to bind to heme proteins such as CytOx and methemoglobin. While H_2S interaction with the former has been implicated in its toxicity, the interaction between methemoglobin and H_2S is a possible detoxification pathway. Smith and Gosselin (1964) pretreated rabbits and armadillos with sodium nitrite, an inducer of

methemoglobin, and reported increased survival following intracardiac and intravenous (i.v.) injection with lethal doses of sodium sulfide. The investigators also reported that female Charles River CD1 mice pretreated with sodium nitrite and *p*-aminopropiophenone, also an inducer of methemoglobin, had increased survival following i.p. injection with sodium sulfide (60 mg/kg). Survival time also significantly increased in mice exposed to 722 and 985 ppm (1003 and 1369 mg/m³) H₂S and pretreated with sodium nitrite and *p*-aminopropiophenone.

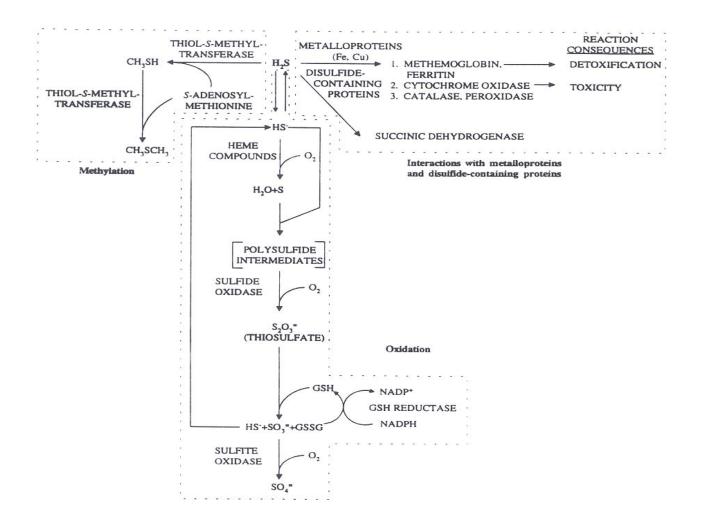


Figure 1. Metabolic pathways of hydrogen sulfide.

Source: adapted from Beauchamp et al., 1984.

3.3. DISTRIBUTION

Forensic autopsies of three men who were overcome by H_2S (estimated concentration 767-906 mg/m³ or 552-652 ppm) demonstrated that the gas is widely distributed in exposed humans (Kimura et al., 1994). Blood sulfide concentrations of the three victims were 0.1, 0.2, and 0.8 µg/g 2-3 hours after death. At 24-hours after death, blood sulfide was measured at 0.5, 0.23 µg/g, and not detected, respectively. Sulfide was also detected in the brain (0.2-1.06 µg/g), lung (0.21-0.68 µg/g), liver (1.30-1.56 µg/g), kidney (0.47-1.50 µg/g), and spleen (0.32-0.62 µg/g) 24 hours after exposure. The investigators state that postmortem generation of H_2S by the blood, liver, and kidney was extremely remarkable making the values from these three victims unreliable. However, sulfide is not normally detected in the brain and lung of control postmortem specimens, and sulfide detected in these tissues is considered a reliable indicator of exposure. Imamura et al. (1996) also reported detectable concentrations of sulfide in the blood, lung, brain, liver, and kidney of a worker overcome by H_2S . The investigators also reported that sulfide was detected in the skeletal muscles of the abdomen and leg.

Distribution of H_2S in animals is very similar to humans. Kohno et al. (1991) exposed male Wistar rats to 75 ppm (104 mg/m³) for 20, 40, or 60 min. Following exposure, animals were sacrificed and blood and tissues (brain, lung, heart, liver, spleen, and kidney) were collected within five min of sacrifice. After 20 minutes of exposure, the concentration of sulfide in the blood, brain, lung, heart, liver, spleen, and kidney was approximately 10, 25, 22, 38, 23, 27, and 30 µg/g of tissue, respectively. The concentration of sulfide in these tissues was relatively constant regardless of the exposure duration.

3.4. ELIMINATION

Urine is the primary route of elimination following H_2S exposure. Gunina (1959) reported that following exposure to sodium sulfide via i.v. and subcutaneous (s.c.) routes or exposure to H_2S by inhalation routes in dogs and rats the majority of the dose (70-99%) was eliminated in the urine by 24 hours postexposure. Kage et al. (1992) detected thiosulfate but not sulfide in the urine of rabbits exposed to 100-200 ppm (139-278 mg/m³) H_2S by inhalation up to 24 hr after exposure with the highest levels being detected 2 hr after exposure. Thiosulfate was also found in the urine of volunteers exposed to 8, 18, or 30 ppm (11, 25, or 42 mg/m³) H_2S for 30-45 minutes and maintenance workers in a pelt processing plant with known exposures to H_2S (Kangas and Savolainen, 1987).

4. HAZARD IDENTIFICATION

4.1. STUDIES IN HUMANS—EPIDEMIOLOGY, CASE REPORTS, CLINICAL CONTROLS

On December 14, 2001, after internal peer review of this document, the Agency articulated its interim policy on the use of third-party studies submitted by regulated entities (U.S. EPA, 2001). For these purposes, EPA is considering "third party studies" as studies that have not been conducted or funded by a federal agency pursuant to regulations that protect human subjects. Under the interim policy, the Agency will not consider or rely on any such human studies (third-party studies involving deliberate exposure of human subjects when used to identify or quantify toxic endpoints such as those submitted to establish a NOAEL or NOEL for systemic toxicity of pesticides) in its regulatory decision making, whether previously or newly submitted. Some of the supporting studies discussed in this Toxicological Review are third-party studies. The studies cited provide data which suggests and inform a public health concern for hydrogen sulfide, but were not designed or used as principal studies in the derivation of any quantitative value for hydrogen sulfide based on NOAELs or LOAELs. The Agency is requesting that the National Academy of Sciences conduct an expeditious review of the complex scientific and ethical issues posed by EPA's possible use of third-party studies that intentionally dose human subjects with toxicants to identify or quantify their effects.

Recent reviews of the health hazards associated with H₂S exposure and subsequent treatment include Milby and Baselt (1999a) and Guidotti (1996). Earlier reviews of the health effects were provided by Glass (1990), Reiffenstein et al. (1992), and Mehlman (1994). Exposure to H₂S has been reported to be an important cause of morbidity and mortality in the workplace (Snyder et al., 1995) and olfactory dysfunction (Hirsch and Zavala, 1999). These reviews indicate that the typical "rotten-egg odor" of H₂S is an inadequate warning indicator of exposure since levels in the range of 100-200 ppm (140-280 mg/m³) can lead to loss of smell followed by olfactory paralysis (Reiffenstein et al., 1992). These authors estimate the odor threshold of H_2S in the range of 3 to 20 ppb (0.004 to 0.03 mg/m³). The case reports cited below involving overexposure have documented a variety of central nervous system (CNS) transitory symptoms, such as dizziness, nausea, headache and others more long-acting such as abrupt physical collapse or "knockdown," all of which have been attributed to direct effects of H₂S on the brain (Milby and Baselt, 1999b). Prolonged unconsciousness can lead to respiratory failure, hypoxia, and death. In the reviews cited above, levels associated with "knockdown" have been estimated to be in the range of 500-1,000 ppm (695-1,390 mg/m³). Levels associated with pulmonary edema have been estimated to be in the range of 250-500 ppm ($348-695 \text{ mg/m}^3$). A rather extensive list of acute symptoms of intoxication resulting from an accidental release of H₂S in an industrial setting was provided by Kleinfeld et al. (1964).

The following case studies describe exposures to a number of volatile and semivolatile compounds in addition to hydrogen sulfide (e.g., liquid manure) whereby the actual toxic agent is unclear. The point of these descriptions is that exposure to high concentrations of hydrogen sulfide, even for very brief periods, can have life-threatening consequences involving respiratory arrest and secondary CNS effects as well as persistent neurological sequelae.

Case Studies:

A 19-year-old oil rig worker was exposed to unspecified concentrations of H_2S (Burnett et al., 1977) rendering him unconscious for an indeterminate amount of time. Upon resuscitation, he exhibited malaise, anterior chest pain, dyspnea, headache, nausea and vomiting, tearing of the eyes and photophobia and coughed up blood. Upon arrival at the hospital for further treatment, his vital signs were normal and he was no longer in respiratory distress. He had severe photophobia and blepharospasms, but no signs of conjunctivitis. He also possessed a cough and some motor weakness of his right arm and leg. A neurologic examination and a chest x-ray revealed no abnormalities. After a 3-day stay in the hospital, he was discharged.

Two fatalities were due to massive aspiration of liquid manure and one fatality was due to severe pulmonary edema with no aspiration of the manure (Osbern and Crapo, 1981). Another 41-year-old individual fell unconscious into liquid manure during a rescue attempt. After resuscitation, he had difficulty breathing and was agitated, but he exhibited no focal neurologic deficits. His initial chest radiograph showed a five-lobed alveolar infiltrate. After two weeks in the hospital, his chest radiograph showed improvement and his lung function was normal except for a slightly reduced maximum midexpiratory flow rate. Six months after the accident, he had a normal chest radiograph and was asymptomatic. Blood sulfide levels of the two individuals who had fatally aspirated manure were 5.0 and 3.6 mg/L. The individual who had died by pulmonary edema had a blood sulfide level of 0.8 mg/L. Control blood samples exhibited sulfide levels below 0.05 mg/L while blood samples from random autopsy cases of badly decomposed subjects did not exceed 0.4 mg/L. No blood sulfide measurements were performed on the 41-year-old survivor. Eight days following the accident, air analyzed from the liquid manure storage tank contained 6,360 ppm methane, 400 ppm carbon monoxide, 1.5 ppm ammonia, 2% carbon dioxide, 18% oxygen, and 76 ppm (106 mg/m³) H₂S.

Another case study illustrates the possible long-term sequelae of H_2S exposure. A 30year-old man displayed dyspnea, chest tightness, and haemoptysis following exposure to a toxic gas in a lavatory facility at his place of work (Parra et al., 1991). The facility was connected to a manure pit; no measurements of H_2S were performed. Physical examination and routine laboratory studies revealed no abnormalities. However, a chest radiograph detected a mild bilateral interstitial pattern. Bronchoscopy showed a widespread reddish mucosa. Pulmonary function tests showed a mild restrictive disease. After five months, the patient possessed residual exertion-dyspnea but was otherwise asymptomatic. The diagnosis was pneumonitis caused by the inhalation of a toxic gas. Other exposed workers exhibited nausea, vomiting, dizziness, dyspnea, and eye and nose irritation. One of the exposed workers died a few hours after exposure. An autopsy revealed haemorrhagic bronchitis and the cause of death was asphyxia attributed to the inhalation of a "toxic gas," with H₂S suspected as a major component.

Several case studies report on the rapid toxicity following exposure to high levels of H_2S . A 14-year-old boy found a discarded cylinder containing H_2S in a rubbish dump and immediately died when he opened the tank (Allyn, 1931). His father also died during a rescue attempt. Both bodies were deeply cyanosed. A fatality, apparently due to release of H_2S from a load of refinery waste was described by Simson and Simpson (1971). This report also mentioned eyewitness accounts of intense central cyanosis in the victim.

At a poultry feather fertilizer plant, a worker was exposed to H_2S while attempting to repair a leak and was killed (Breysse, 1961). Hydrogen sulfide was created as a byproduct of the putrefaction of the feathers, and was eliminated through a pipe leading to a sawmill log pond where it was discarded. In the lungs of the victim, the alveolar spaces were filled with edema fluid and numerous pigment-filled macrophages. The diagnosis was pulmonary edema, and the cause of death was H_2S inhalation. Measurements of H_2S concentrations at various locations in the fertilizer plant revealed levels as high as 4,000 ppm (5,560 mg/m³) during the cooking and putrefaction of feathers.

A 16-year-old boy suffered fatal H_2S exposure during transport of liquid manure (Hagley and South, 1983). He was found at the bottom of the manure tank, pale, unconscious and apneic. There was no evidence that he had aspirated any manure. The boy began breathing following resuscitation and his color returned to normal. His heart rate and blood pressure were normal though he did not respond to painful stimuli. Over the following hour, he became responsive to stimuli, but then developed extensor spasms and began to hyperventilate. A chest radiograph and a CT scan of the head were normal. The patient developed pneumonia. However, his neurological condition deteriorated, and he suffered partial seizures and exhibited a decerebrate response to painful stimuli. The patient died five days after the accident with clinical signs of brain stem damage. Postmortem examination revealed that he had cerebral edema. A week later, H_2S was measured 30 cm below the manhole of the tank and was found to exceed the upper limit of detection of the monitoring equipment at 150 ppm (208 mg/m³).

By contrast, Milby (1962) has reported two cases of severe, but non-fatal H_2S poisonings. While disposing of a H_2S -filled gas cylinder and the subsequent release of its contents, two men that were exposed to the gas collapsed immediately and underwent convulsions while unconscious. Upon being rescued from the gas, the men required artificial respiration to begin breathing on their own. Both men were hospitalized in an unconscious state and were later revived. At the conclusion of their hospital stay, both men were without symptoms. The author states that they "remained well to the present," but it is unclear how long an interval that represents.

A mass exposure to H₂S took place among workers laying the foundation of a municipal sewage pumping station, leading to the death of a police officer attempting to rescue an unconscious worker and persistent neurological sequelae in at least one other individual (Snyder et al., 1995). Complications included: decreased ability to communicate, slow speech, marked visual memory impairment with poor acquisition, and difficulty retaining and recalling new information. The findings were reported as essentially unchanged 12-18 months after exposure. Physical findings in affected individuals were limited to pharyngeal and conjunctival erythema and corneal abrasions. Additional reports associated H₂S exposure (via sewer gas) with lung function and neurological decrements and, in certain studies, deaths were reported by Watt et al. (1997) and Hall and Rumack (1997).

Tvedt et al. (1991a) described the follow-up of six H_2S -exposed patients who lost consciousness and experienced signs of long-term neurological sequelae associated with hypoxic brain damage. Immediate symptoms included cyanosis, pulmonary edema, seizures, and coma. Delayed symptoms (5-10 year follow-up) ranged from worker disability due to neurological symptoms to brain damage severe enough to qualify as dementia in one individual. Five patients who had been unconscious in an H_2S atmosphere for 5-20 minutes showed persistent impairment with memory and motor function most affected. The two patients with the most serious symptoms developed pulmonary edema. No measures of H_2S in air, blood or urine were reported. The investigators concluded that the outcome of acute exposures can be variable and that neurological symptoms may not become apparent until a considerable period of time has elapsed.

Tvedt et al. (1991b) also described delayed neuropsychiatric symptoms in an individual who became unconscious for 15-20 minutes during an H_2S exposure. He regained consciousness after two days, but then deteriorated again and remained comatose for a month. Upon waking up, the individual was amnesiac, nearly blind, had reduced hearing and ataxia. At follow-up (five years later) he had not been able to resume work and evidenced slight motor, memory and visual impairments and clinical indications of slight cerebral atrophy. Electroencephalogram and evoked responses were normal at this time.

Symptoms similar to those described by Tvedt et al. (1991b) were also reported by Kilburn (1993) in an oil well tester rendered semiconscious from exposure to high levels of H_2S . Follow-up (39 and 49 months) symptoms suggested damage to the brain stem, basal ganglia, vestibular apparatus, cortex, and other brain structures. There was profound impairment of cognitive function, memory, visual perception and coordination, intelligence and neurophysiologic functions. Malingering was ruled out by the reproducibility of performance during testing sessions.

Persistent neurocognitive impairment from H_2S poisoning was also reported by Wasch et al. (1989). Three cases are described with 2/3 individuals experiencing unconsciousness during exposure. All three individuals had moderate to severe attention and concentration deficits, as well as varying degrees and types of memory, learning, and perceptual deficits and prolonged abnormal event-related potentials (P-300 latency). In one individual, symptoms persisted at follow-up three years after exposure.

In another study, a 20-month-old child developed intermittent paroxysmal tonic deviation of the eyes (Gaitonde et al., 1987). After a few months the abnormal eye movements resolved, progressive involuntary movements of the entire body developed, and the child fell frequently. The child was admitted to the hospital with gross truncal ataxia, choreoathetosis, dystonia, and an inability to stand. The child was dysarthic but had normal eye movements. Computer tomography revealed bilateral areas of low attenuation in the basal ganglia and some of the surrounding white matter. Virology was negative and there was no evidence of streptococcal infection. The brain scan suggested toxic encephalopathy. Shortly after admission, the child's condition improved spontaneously. Ten weeks after admission, the ataxia had resolved and choreathetoid movements were reduced. A repeat brain scan was normal. The child's illness was attributed to H₂S exposure. The family lived next to a coal mine where a burning tip emitted H₂S for nearly one year. Three months prior to the child's admission, H₂S emissions were monitored for four months, and the maximum recorded level in family's housing scheme was 0.6 ppm (0.8 mg/m^3). Local authorities admitted that H₂S levels may have been many times higher prior to monitoring. The burning tip had been extinguished just prior to the child's admission to the hospital. This study is discussed at length in Section 4.7.1., Possible Childhood Susceptibility.

Epidemiological Studies:

Chronic low-level health effects in two communities exposed to industrial sources of H_2S were assessed by means of a symptom-survey using trained interviewers (Legator et al., 2001). Results of the survey were compared to health effects in three reference communities not exposed to H_2S . Symptoms were elevated in both exposed communities compared to the referent populations in: CNS effects, respiratory effects, and blood (e.g., anemia, clotting, etc.). Although the odds ratios were elevated over controls, they fell within the confidence intervals. In addition, respondents from the two exposed communities had long-standing involvement with H_2S exposure issues which may have contributed to study bias. Subjects in one of the exposed communities were also assessed by means of several objective psychoneurological tests, the results of which were consistent with results of the symptom survey (Borda, 1997; cited by Legator et al., 2001).

A mortality study was undertaken among Finnish sulfate mill workers exposed to H_2S and organic sulfides (Jappinen and Tola, 1990). Workers had been employed for at least one

year between 1945 and 1961 at three pulp and paper mills owned by the same company. No exposure data were presented. Deaths from all causes were not increased. However, workers exposed to H_2S and organic sulfides exhibited an increase in cardiovascular-related deaths compared to national death rates (37 observed compared to 24.7 expected). Cardiovascular mortality was higher in workers employed for ≥ 5 years compared to workers exposed for 1-4 years. The investigators state that increased mortality could not be explained by common risk factors and that differences in smoking habits did not explain the findings. They suggest that increased mortality may have been associated with H_2S and organic sulfide exposure.

A cross-sectional study investigated the effects of presumed H_2S exposure in sewer workers to determine if chronic exposure to the gas was associated with decreased lung function (Richardson, 1995). Sixty-eight sewer workers performed spirometric tests and results were compared to 60 nonexposed water treatment workers. Job titles were used to categorize sewer workers according to presumed H_2S exposure levels. There was a statistically significant decrease in mean FEV₁/FVC in sewer workers compared to water treatment workers. The effect was greater in sewer workers presumed to have higher exposures to H_2S and longer exposure histories. In nonsmoking subjects, sewer workers were only able to attain 89% of the predicted FEV₁/FVC value compared to 98% in water treatment workers. Although the study author states that chronic low level exposure to H_2S may be associated with decreased lung function, no measurements of H_2S were made so any quantitative relationship between "low levels" of exposure and effect are speculative.

The pulmonary effects resulting from H_2S exposure were assessed in 175 workers who extracted and processed oil and gas (Hessel et al., 1997). The workers received a questionnaire concerning sour gas exposures that caused symptoms or loss of consciousness. Thirty-four percent of the respondents indicated that they had exposures serious enough to cause symptoms, and 8% of the workers stated they had experienced a loss of consciousness due to sour gas exposure. In workers that experienced symptoms, no decrease in spirometric values or excess symptoms were noted. While spirometric values were also not affected in workers that lost consciousness, the workers experienced shortness of breath during physical activity, wheezing with tightness in the chest, and wheezing attacks. The investigators stated that these symptoms are consistent with bronchial hyperresponsiveness. The data also demonstrate the refractiveness of the lung to the effects of H_2S , with lung function parameters being unchanged despite unconsciousness from H_2S exposure.

A cross-sectional pulmonary function study was performed in males that worked in viscose rayon plants (Higashi et al., 1983). A one-workday study was performed in which workers were monitored for H_2S exposure for 8 hr and had a forced expiratory flow-volume test performed. Hydrogen sulfide levels were determined using passive diffusion dosimeters worn by the workers (n = 30 in exposed and matched control workers). The workers had been

exposed to H_2S for an average of 12.3 years. Occupational exposure levels in the exposed workers ranged between 0.3 and 7.8 ppm (2.9 ppm average; 4 mg/m³) compared with < 0.1 ppm (0.1 mg/m³) in matched control workers. No significant differences in the results from the pulmonary function tests were observed in exposed compared to nonexposed workers. Also, no significant differences were noted in the pulmonary function tests taken at the beginning and ending of work shifts.

In addition to the one-workday study, a cross-sectional study was undertaken on 115 exposed workers and 209 nonexposed workers. The subjects underwent forced expiratory flow-volume tests, but were not monitored for exposure. The investigators (Higashi et al., 1983) concluded that chronic exposure to low levels of H_2S would not cause adverse pulmonary health effects.

A survey of 123 male viscose rayon workers exposed to H_2S and/or carbon disulfide (CS_2) and 67 nonexposed metal, plastic, and starch workers was conducted in which subjects were questioned about eye irritation complaints (Vanhoorne et al., 1995). Exposed and control workers were employed for at least one year by their employer. Exposure to H_2S and CS_2 was determined using personal monitoring pumps. Exposure to H_2S ranged between 0.2 and 8.9 mg/m³. Workers exposed to H_2S also tended to have high exposure to CS_2 . No workers were exposed to H_2S only. Several types of eye complaints were reported by workers exposed to CS_2 and H_2S including: pain, tension, burning, hazy sight, photophobia, and irritation. While the percentage of complaints increased with exposure, the frequency of complaints was higher in workers exposed to H_2S and CS_2 compared to workers exposed to CS_2 alone. The investigators conclude, however, that deciding which of the two gases was responsible for eye irritation could not be determined, but argue that *a priori* of evidence suggests that H_2S is responsible. Considering all of the evidence, the investigators further suggested that CS_2 enhances the eye irritation properties of H_2S .

A study by Kilburn and Warshaw (1995) included 13 former workers and 22 near-plant residents of a desulfurization unit within a refinery processing "sour' crude oil. Air monitoring for H_2S and other chemicals at street level near the subjects' homes during one week in July 1990 revealed that the H_2S concentration was 0.01 ppm with periodic peaks of 0.1 ppm along with mercaptans (0.002 ppm), ethane (0.5 ppm), propane (0.5 ppm) and unquantified levels of vanadium and thiodiglycolic acid. In addition, dimethylsulfide, and vanadium (as vanadium pentoxide) were detected in air. Reduced sulfur gases measured outside the facility showed 24-hr averages of up to 21 ppm mercaptans, 0 to 8.8 ppm H_2S , carbon oxide sulfide (COS: 6 to 52 ppm) and for total reduced sulfur 6 to 71 ppm.

Questionnaires were completed by subjects and controls. Neurobehavioral functions and a profile of mood states were compared to 32 controls, matched by age and educational level. Neuropsychological functions were also evaluated, including: two-choice reaction time, balance,

color discrimination, digit symbol, trail-making A and B, and immediate recall. Abnormal mean values of exposed subjects compared to controls were determined to be statistically significant for these functions. Mood state scores were found to be much higher in the exposed subjects than the controls. However, many of the neurophysiological findings had large standard deviations and consequently were highly-variable. While alcohol use was excluded as a confounder, all of the ex-workers had been exposed to solvents. Although the authors concluded that long-term exposure to "low doses" of H₂S is a plausible cause of the neurological function deficits, they acknowledged that other effects were attributed to sulfur dioxide and vanadium pentoxide). Given the complexity of the exposure environment - in terms of the numbers of chemicals monitored, the few numbers of individuals studied and the very limited exposure measurements - it is premature to attribute the deficits cited with H₂S exposure.

Neurobehavioral effects of H₂S were studied in 16 individuals (male and female, ages 21 to 68 years) 2 to 22 years after exposure (Kilburn, 1997). Five individuals had been unconscious after exposure and four had downwind exposures. The exposure level was not reported in any of the cases. Five of the 16 individuals were categorized in the "chronic" exposure group, exposed for 11-22 years. Five were in the "acute" exposure category having been exposed for minutes and the remaining six were in the "hours" category. Twelve of the 16 individuals were plaintiffs in lawsuits. Results from a battery of tests were compared to those from 353 national referents, matched by age, sex, and years of education. Smoking status was not a significant factor. Neurophysiologic and neuropsychologic tests, the latter adjusted for educational level, were employed. Permanent neurobehavioral impairment was apparent in all 16 subjects. Brief high doses were "devastating," whereas protracted low doses of individuals in the chronic category showed effects (p<0.05) on the more sensitive neurobehavioral tests, such as balance, visual fields, and choice reaction time. There was no indication of bias from the lawsuit and it was considered unlikely that the symptoms could be attributed to other causes. Kilburn (1997) concluded that moderate occupational exposure and downwind environmental exposure can cause permanent impairment. This finding was assessed in a subsequent study.

Neurobehavioral deficits attributed to low-level environmental exposure of 103 individuals to H_2S were reported by Kilburn (1999). The neurobehavioral battery included 28 tests. One group of 24 homeowners was exposed to H_2S that had collected in crawlspaces and under concrete foundations. Concentrations ranged from 0.1 to 1.0 ppm with several peaks up to 5 ppm during a two-month period in 1996. The homeowners were assessed about one year later. They exhibited abnormal balance, color discrimination, grip strength and delayed verbal recall. It was not stated if the exposure was ongoing at the time of the neurobehavioral assessment. In another group, 48 adults were exposed to measured peak levels ranging from 1 to 20 ppm at street level during a refinery explosion and fire in 1992. They were assessed for neurobehavioral deficits four years later in 1996. The subjects had 11-12 abnormalities compared to 4 in the first group. A third group consisted of 16 patients with a wide range of exposure who were assessed after latent periods of 1.7 to 22 years [these were the same individuals discussed in the preceding paragraph by Kilburn (1997)]. A fourth group consisted of 13 workers and 22 residents living at or near stack gas emissions from a refinery and desulfurization plant. These individuals were also exposed carbon oxide sulfide (2.6 - 52.1 ppm), mercaptans (1.0 - 21.1 ppm), and 24-hr levels of H₂S at 0 to 8.8 ppm. Referents (357) were selected from four towns (in four states) free of known chemical contamination. Analysis of variance and covariance adjusted for differences in age, education, gender, and height. Kilburn (1997) concluded that peak dose rather than duration of exposure is predictive of effects, and those exposed to nonlethal levels do not recover completely from H₂S. Given the small number (24) of homeowners in the first group exposed to low levels and the type of deficits observed, it is premature to ascribe cause and effect to H₂S.

A five-year retrospective study (Arnold et al., 1985) of 250 workers exposed to H_2S considered compensation claims submitted from 1979-1983. Fifty-four percent of the workers became unconscious after exposure. Of the clinical findings, neurologic, respiratory and opthalmologic ailments predominated. The overall fatality rate was 2.8%.

In conclusion, H_2S exposure levels that result in temporary unconsciousness (e.g., 15-30 minutes) can cause profound neurophysiological, neurobehavioral, neurocognitive, neurophysical, respiratory, and opthalmologic deficits that are persistent. While some of these deficits can occur when the exposed individual remains conscious, it is less clearly established if intermittent or continuous exposure to "low levels" of H_2S cause similar persistent symptoms.

4.2. PRECHRONIC AND CHRONIC STUDIES IN ANIMALS

4.2.1. Oral

Wetterau et al. (1964) examined the effects of H_2S in chickens and pigs that were fed dried green animal fodder. The fodder had been dried by sulfur-containing brown coal or other fuels containing H_2S . The fodder was analyzed for H_2S content by acidification with hydrochloric acid and iodometric determination prior to animal administration. The H_2S content in all of the experiments ranged from 0.035-0.121%. In chickens (50 per group), no adverse effects were observed in feed intake, body weight, or survival in animals fed H_2S -treated alfalfa (2-12% of total feed) and H_2S -treated wheat bran (1-12% of total feed) daily for up to 70 days. Adult pigs (numbers not given) were fed diets containing 100, 200, or 400 grams of H_2S -treated dried alfalfa (4, 8, and 24% of the diet, respectively) daily for 105 days. Pigs fed 200 or 400 grams of H_2S -treated dried alfalfa had decreased weight gains of 4 and 22%, respectively, compared with the control group. No statistical analysis was presented. Food intake decreased by 33% in animals fed 400 grams of H_2S -treated dried feed per day. Diarrhea was reported in adult animals that consumed diets containing no treated dried feed that was suddenly changed to diets containing 24% treated dried feed.

In a separate experiment, weaned pigs (numbers not provide) were fed diets containing 100, 200, or 400 grams of H_2S -treated dried alfalfa (4, 8, and 24% of the diet, respectively) daily for 105 days. In weaned animals previously exposed to treated dried feed, the animals did not develop diarrhea when the amount of treated dried feed was increased to 24% of total feed intake. The percent of H_2S in the diet containing 24% dried feed was 0.121%. The investigators concluded that there was no indication that diets containing a high amount of H_2S could cause adverse health effects in animals. The diarrhea observed in naive adult pigs may have been caused by the sudden change in diet rather than the presence of H_2S , since increasing the feed containing H_2S in weaned pigs previously exposed to a similar diet had no adverse effects. Decreased weight gain observed in the second study is apparently due to a decrease in feed intake. The pigs in this study did not develop diarrhea, but there was a decrease in food intake. The investigators reported that previous data suggested that animals refused to eat food containing H_2S , which may explain the decreased feed intake and the subsequent decreased weight gain observed in the study. Results of the study were likely caused by food avoidance from odor rather than toxicity.

A study by Wetterau et al. (1964) was used previously on IRIS as the basis for deriving an RfD. However, the low concentration of H_2S in the feed, estimated at only about 15 mg/kgday, coupled with the investigators' conclusions that the diets did not cause adverse or reproducible health effects in pigs precludes the use of this study in the derivation of an RfD in this current assessment.

Male and female Sprague-Dawley rats (20/sex/dose group) were administered 0, 1, 3.5, or 7 mg/kg-day H₂S by gavage daily for 89 days (Anderson, 1987). Dose-levels were determined after a 14-day range-finding study. Hydrogen sulfide doses were prepared by purging deoxygenated deionized water with pure H₂S gas. Prepared H₂S solutions were stored in amber vials with Teflon seals and fresh vials were used at each dose administration. Dosing solutions were prepared weekly and solution stability was determined. Animals were examined for mortality at the beginning and end of each working day, and clinical signs were monitored immediately after dose administration. Food consumption of 10 rats per dose group was determined weekly. Animals were weighed weekly during the treatment period and at necropsy. Blood samples were collected from the suborbital sinus of animals prior to study initiation, days 27 to 31 of treatment, and before necropsy. Blood and platelet counts were determined, and clinical chemistry was performed. At study termination, animals were sacrificed by carbon dioxide asphyxiation. Sacrificed animals and animals that died during the treatment period were subjected to a full necropsy. In control and high-dose animals, liver, kidney, spleen, gonad,

brain, heart, adrenal, and gross lesions were examined microscopically. In the low- and mid-dose groups, only lung, kidney, liver, and gross lesions were examined microscopically. Mortality was 50% in the high-dose males compared to 5% in control males. No deaths were observed in high-dose females. Mortality in mid-dose males and females was 10 and 5%, respectively, and no deaths were observed in low-dose animals or control females. Weekly mortality is reported in Table 1.

		Weeks													
	Dose (mg/kg-day)	1	2	3	4	5	6	7	8	9	10	11	12	13	%
Males	0	0	0	0	1	1	1	1	1	1	1	1	1	1	5
	1	0	0	0	0	0	0	0	0	0	0	0	0	0	0
	3.5	0	0	0	0	1	1	1	1	1	1	1	1	2	10
	7	0	0	0	0	2	2	4	4	6	6	7	9	10	50
Females	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
	1	0	0	0	0	0	0	0	0	0	0	0	0	0	0
	3.5	0	0	0	0	0	0	0	1	1	1	1	1	1	5
	7	0	0	0	0	0	0	0	0	0	0	0	0	0	0

Table 1. Cumulative weekly mortality in male and female Sprague-Dawley rats administered H_2S orally

Source: Anderson, 1987

Compound-related neuromuscular and behavioral signs were observed in high-dose male and female animals, with the males displaying a higher incidence of these effects (Anderson, 1987). High dose-males exhibited abnormal posture, convulsions, fist clutching, sedation, unusual vocalization, piloerection, labored breathing, and increased respiratory rates. High-dose females exhibited convulsion, sedation, and labored respiration. Although restlessness was observed in all treated males, and increased salivation in all treated females, these clinical signs generally occurred sporadically, and the incidence of these findings did not show a clear doseresponse. Therefore, the toxicological significance of restlessness in treated males, and salivation in treated females, is not clear.

No significant differences in body weights were noted in any of the treated animals (Anderson, 1987). In addition, no differences in food consumption were noted between treated and control groups. There were no treatment-related effects in hematology or clinical chemistry.

Ophthalmological findings prior to sacrifice were normal. The most common findings at necropsy were pink and red mottled lungs, which occurred in all groups. In high-dose males that died prior to sacrifice, pink-red discoloration of the lungs, nasal discharge, penile discharge, reddening of the scrotal sac, and darkening of the spleen were noted. Although absolute organ weights were not affected by treatment, relative heart weights increased (p < 0.05) in high-dose males, and relative kidney weights increased (p < 0.05) in high-dose females. No gross lesions could be attributed to compound administration and all microscopic findings were considered incidental to compound administration.

It should be noted that there are a number of problems with this study that prevent its use in the derivation of an RfD. For example, a 75 to 100% incidence of pneumonia was reported in the vehicle control and treated groups, however, a 0% incidence of pneumonia was reported for the nongavaged, rack control group. This information indicates procedural problems with the dosing of the animals. Also, there was one group where mortality was 50%, the high dose male group. However, the cause of death was not determined. In addition, a malady diagnosed as ulcerative dermatitis was common across all groups. Good Laboratory Practice would preclude the acceptance of data from such diseased animals for drawing conclusions regarding the effects of H_2S .

4.2.2. Inhalation

The Chemical Industry Institute of Toxicology (CIIT, 1983a) performed a 90-day inhalation toxicity study in B6C3F1/CrlBr mice using H₂S. Animals were individually housed and exposed to 0, 10.1, 30.5, or 80 ppm (0, 14, 42, or 111 mg/m³) H₂S for 6 hr/day, 5 days/week for 90 days (10 mice/sex/group). Animals were observed twice daily for mortality and clinical signs. Body weight and feed consumption were determined weekly beginning just before the first exposure. A final fasted body weight was determined for all animals just before necropsy. Each animal also received an ophthalmological exam before the first day of exposure and within seven days of necropsy. A neurological exam was performed before necropsy in which posture, gait, facial muscle tone, and pupillary, palpebral, extensor thrust, and crossed-extensor thrust reflexes were assessed. Animals were housed in metabolism cages for 12 hr prior to necropsy. Urine samples were collected for all animals and volume, appearance, occult blood, specific gravity, protein, pH, ketone, and glucose were determined. On the day of necropsy, animals were anesthetized with ether and blood was drawn from the suborbital sinus (hematology) and abdominal aorta (chemistry). Animals were sacrificed on study day 91, 92, 93, 94, 95, and 99. The brain, kidney, spleen, liver, heart, and ovaries/testes were removed, weighed, and examined for gross and histopathologically observable abnormalities. In addition, the following tissues were examined microscopically: cerebellum, cerebrum, medulla, optic nerve, spinal cord, sciatic and anterior tibial nerves, eyes, pituitary, thyroid, parathyroid, salivary glands, heart, lungs,

spleen, liver, pancreas, adrenals, mesenteric and mandibular lymph nodes, kidneys, bladder, lacrimal glands, ovaries, uterus, oviducts, vagina, cervix, stomach, duodenum, ileum, jejunum, large and small colon, caecum, skeletal muscle, skin, mammary glands (both sexes), femur, bone marrow, aorta, ear canal with zymbal gland, nasal turbinates (four levels), trachea, testes, epididymis, esophagus, thymus, prostate, seminal vesicle, and gross lesions.

Body weights were significantly depressed in both males and females in the highexposure groups with sporadic significant decreases in the lower male exposure groups (CIIT, 1983a). Decreases in body weights tended to be more pronounced in the male and female high exposure groups beginning in the 6th week of exposure. Feed consumption was significantly reduced in the high-exposure animals. There were no toxicologically- or statistically-significant differences in hematology, serum chemistry, urinalysis, ophthalmology, or neurological function except for a 30.5 ppm (42 mg/m³) female and 80 ppm (111 mg/m³) male that did not respond to artificial light stimulus.

Two female controls and one 30.5 ppm (42 mg/m³) male died prior to study termination. Both control animals exhibited multiple focal black discoloration of the glandular stomach. The 30.5 ppm (42 mg/m³) male exhibited a skull fracture with red discolorations and trauma apparently due to a feeding accident. A female and male in the 80 ppm (111 mg/m³) exposure group were sacrificed *in extremis* on study days five and six, respectively. Both sacrificed animals exhibited prostration and/or hypoactivity prior to sacrifice. Both animals also had the same black discoloration of the glandular stomach as those in the control group. A low incidence of alopecia and emaciation was observed in control and exposed animals. In addition, a 10.1 ppm (14 mg/m³) male lost the use of anterior appendage with the appearance of paralysis while one 30.5 ppm (42 mg/m³) female and a 80 ppm (111 mg/m³) female were observed to have a missing front appendage. The investigators stated that it was unknown how the male lost its appendage but that the female's loss was apparently congenital. The investigators considered none of the above findings to be compound-related. Gross pathology of surviving animals also

The mean absolute weights of the heart, liver, and spleen from male mice exposed to 80 ppm (111 mg/m³) H₂S were significantly depressed. In addition, the mean absolute kidney weight of 80 ppm (111 mg/m³) female mice was also significantly reduced by approximately 20%. However, relative weights were not significantly different and clinical pathology and histology were negative in both sexes. Histological examination of surviving animals revealed the nasal tract as the only site where lesions were considered compound-related. Male (8/9) and female (7/9) mice exposed to 80 ppm (111 mg/m³) H₂S exhibited minimal to mild inflammation of the anterior portion of the nasal mucosa (section I; section I and II in two mice). The lesion was primarily located in the squamous portion of the nasal mucosa but extended into the respiratory type epithelium (ciliated) in some animals. In one female, the lesion was suppurative

and severe involving the entire nasal passage and associated structures. The lesion was also observed in two 80 ppm (111 mg/m³) mice that were exposed *in extremis*. No other histological findings were considered compound-related (CIIT, 1983a).

The critical effect in mice was inflammation of the nasal mucosa. This effect was present in male (8/9) and female (7/9) animals exposed to 80 ppm (111 mg/m³) H₂S. The lesions, judged as minimal to mild in severity, were localized primarily in the squamous portion of the nasal mucosa but extended to regions of ciliated respiratory-type epithelium. Since H₂S is an irritant gas and other researchers (Brenneman et al., 2000; Dorman et al., 2000; Lopez et al., 1987) have reported nasal inflammation following H₂S exposure, the critical effect in mice was considered to be inflammation of the nasal mucosa. The LOAEL for this effect in mice was 80 ppm (111 mg/m³) and the NOAEL was 30.5 ppm (42 mg/m³) (CIIT, 1983a).

With the following exception, the same methods described in the CIIT (1983a) mouse study were used in a similar study using Fischer 344 rats (CIIT, 1983b). Following a 90-day exposure period in rats, 10 males and females from each group were selected for clinical pathology and histology, while the remaining 5 males and females from each group were used for special neuropathologic studies. In the neuropathologic studies, rats were anesthetized with sodium pentobarbital containing 200 units of heparin. Rats were perfused via the left ventricle with a 4% phosphate buffered glutaraldehyde solution. The perfused animals were then refrigerated at 4°C overnight. On the following day, the right and left sciatic nerves and the cervical and lumbar portions of the spinal cord were dissected and placed in 4% glutaraldehyde. The left sural nerve and large muscle branch of the tibial nerve were oscimated and placed in cedarwood oil for two weeks. Nerve fibers in the oil were teased and individual fibers mounted to glass slides. After staining, specimens were examined by light microscopy for pathological changes.

In Fischer 344 rats, no mortality was observed during the 90-day study (CIIT, 1983b). Body weights were significantly reduced from weeks 1 to 13 in males and females exposed to 80 ppm (111 mg/m³) H₂S. Feed consumption was also significantly depressed in males and females exposed to 80 ppm (111 mg/m³) H₂S. No abnormalities in ophthalmology, neurological function, serum chemistry, or urinalysis were reported. Sulfhemoglobin levels were significantly increased in males exposed to 80 ppm (111 mg/m³) H₂S. Although elevated sulfhemoglobin levels were also observed in females exposed to 30.5 or 80 ppm (42 mg/m³ or 111 mg/m³) H₂S, the levels were not statistically significant. The investigators state that assigning biological significance to the sulfhemoglobin values should be done with caution because of sensitivity of the method and the size of the groups investigated. Also, there were no H₂S-related lesions in animals selected for special neuropathologic studies. Gross pathology revealed a low incidence of lesions that were not compound-related. Organ weights were statistically similar between control and exposure weights except for the relative brain weight of males exposed to 80 ppm (111 mg/m³) H₂S. Relative, but not absolute, brain weights were significantly increased in these animals. Minimal multifocal peribronchial lymphocytic infiltrate was observed in the lungs of all rats. Minimal to mild mononuclear cellular infiltration of sections III and IV of the nasal mucosa and nasolacrimal duct was observed in all rats including the controls. There was also minimal to moderate inflammation of sections I and II of the nasal mucosa, but this effect was not dependent on dose with 45, 40, 75, and 55% of animals affected in the 0, 10.1 (14 mg/m³), 30.5 (42 mg/m³) and 80 ppm (111 mg/m³) exposure groups, respectively. None of the histopathologic changes were considered treatment-related by the investigators. The critical effect was depressed body weight gain in female rats. The LOAEL for this effect was 80 ppm (111 mg/m³) and the NOAEL was 30.5 ppm (42 mg/m³).

The same methods described above (CIIT, 1983a,b), including the special neuropathologic studies in CIIT (1983b) were used in the following study which used Sprague-Dawley rats (CIIT, 1983c). In the Sprague-Dawley rats, 15 males and 15 females per group, no mortality was observed during the 90-day study. Statistically significant decreases in body weights were observed in males exposed to 80 ppm (111 mg/m³) H₂S during weeks 1 through 3, and in 80 ppm-treated females exposed throughout the 90-day study. Feed consumption was also significantly depressed in males and females exposed to 80 ppm (111 mg/m³) H₂S. No abnormalities in ophthalmology, neurological function, hematology, serum chemistry, or urinalysis were reported, and there were no H₂S-related lesions in animals selected for the special neuropathologic studies. Gross pathology revealed a low incidence of lesions that were not compound-related. The black discoloration noted in the glandular stomach of mice was also present in 5 treated rats. However, this effect was not dose-dependent. Organ weights were statistically similar between control and exposure weights, excluding the absolute brain weight of males exposed to 80 ppm (111 mg/m³) H_2S . Absolute, but not relative, brain weights were significantly reduced in these animals; a finding concordant with the results of the CIIT (1983b) study with Fisher 344 rats. The investigators also suggested that the decreased brain weight was due to the chemical nature of H₂S, e.g., low molecular weight and high degree of lipophilicity allowing passage through the blood-brain barrier. Minimal to mild multifocal peribronchial lymphocytic infiltrate was observed in the lungs of all rats including controls. Minimal to mild mononuclear cellular infiltration in sections III and IV of the nasal mucosa and nasolacrimal duct was observed in all rats, including the controls. Although there was also minimal to moderate inflammation in sections I and II of the nasal mucosa, the effect was not related to dose with 35, 35, 30, and 50% of animals affected in the 0, 10.1 (14 mg/m³), 30.5 (42 mg/m³) and 80 ppm (111 mg/m^3) exposure groups, respectively. None of the histopathologic changes were considered treatment-related by the investigators. The critical effect was decreased absolute brain weight in males. The LOAEL for this effect was 80 ppm (111 mg/m³) and the NOAEL was 30.5 ppm (42 mg/m^3).

The following study was actually performed in 1943 by Haskell Laboratory and submitted by Dupont to the Office of Pollution Prevention and Toxics, U.S. EPA (Haskell Laboratory, 1994). Four dogs (sex and species not reported) were exposed to 15 ppm (21 mg/m^3) H₂S 6 hr/day, 5 days/week for 7 weeks. At the start of the week 8, dogs were coexposed to 15 ppm (20.8 mg/m³) H_2S and 10 ppm carbon disulfide. During week 9, dogs were only exposed to 15 ppm (20.8 mg/m³) H_2S which was increased to 30 ppm (42 mg/m³) in weeks 10-12. Blood pressure and heart rate were monitored in the morning and afternoon of exposure days. In addition, blood morphology (not described by the investigators), and arterial and venous CO₂ and O₂ were determined. The frequency with which the investigators determined these parameters was described as "time to time." All measured endpoints were compared to control values that were established during a 4-week preexposure period. The dogs exhibited normal health and behavior during the exposure period. Appetites and weights were maintained satisfactorily, and no blood abnormalities were noted. Abnormal pulse and blood pressure determinations increased in frequency during the course of the experiment (weeks 1-7). The investigators did not state what constituted an abnormal blood pressure reading. However, the investigators suggested a cumulative effect from low exposure. Coexposure with carbon disulfide increased the number of abnormal blood pressure readings. Increasing the concentration to 30 ppm (42 mg/m³) H₂S doubled the number of abnormal blood pressure examinations when compared to a two-week 15 ppm (21 mg/m^3) exposure.

Brenneman et al. (2000) exposed 10-week-old male Sprague-Dawley CD rats (12/exposure group) to 0, 10, 30, or 80 ppm (0, 14, 42.7, or 111 mg/m³) H₂S for 6 hr/day, 7 days/week for 10 weeks. At the end of the 10-week exposure period, animals were euthanized with CO_2 and the noses of the animals were dissected free. The nasal cavities were sectioned at 6 different levels such that sections 3-6 each allowed for histological evaluation of transitional /respiratory and olfactory epithelium. The lesions noted were limited to olfactory tissue, i.e., olfactory neuron loss and basal cell hyperplasia at the olfactory mucosa, and were graded in severity by a subjective scale where 0 = normal, 1 = mild, 2 = moderate, 3 = marked, and 4 = severe.

Table 2 shows that no effects were observed in the control animals or in animals exposed to 10 ppm (14 mg/m³) H_2S (Brenneman et al., 2000). Lesions, limited to the olfactory mucosa, were observed in animals exposed to either 30 (427 mg/m³) or 80 ppm (111 mg/m³) H_2S . These olfactory lesions consisted of multifocal, bilaterally-symmetrical olfactory neuron loss and basal cell hyperplasia affecting the lining of the dorsal medial meatus and dorsal and medial region of the ethmoid recess. The incidence, mean severity, and distribution of the exposure-related lesions increased in a concentration-dependent manner.

		()	10 p	pm	30 p	pm	80 ppm	
Lesion	Nasal Cavity Level	Inc	Sev	Inc	Sev	Inc	Sev	Inc	Sev ^a
Olfactory	3	0/9		0/8		0/6		8/8	2.4
Neuron Loss	4	0/12	_	0/12	_	11/12	1.4	12/12	2.4
	5	0/12		0/12	—	9/12	1.1	11/12	1.5
	6	0/12	_	0/12	_	0/12	_	5/12	1.2
Basal Cell	3 ^b	_	_	_	_	_	_	_	—
Hyperplasia	4	0/12	_	0/12	_	10/12	1.8	12/12	1.2
	5	0/12	_	0/12	_	7/12	1.3	11/12	1.3
	6	0/12	_	0/12	_	0/12	_	6/12	1.0

Table 2. Incidence (Inc) and average severity (Sev) of nasal lesions in male CD rats exposed to $\rm H_2S$

^aAverage severity of the lesion in affected rats in the exposure group: 1 = mild, 2 = moderate, 3 = severe.

^bRats were not evaluated at level 3 because of the common occurrence of basal cell hyperplasia at background levels.

Source: adapted from Brenneman et al. (2000).

Respiratory/transition epithelium largely comprises level 3 of the nose, the olfactory mucosa being only a minor component limited to the most caudal portion of the section. Olfactory neuron loss was only observed in 80 ppm (111 mg/m³) exposure animals at this level of the nose. Basal cell hyperplasia was not evaluated due to the common background occurrence of this lesion at this level.

Olfactory mucosa comprises a slightly larger portion of level 4 of the nasal cavity. This level had lesion distribution restricted to the olfactory tissue in the dorsal medial meatus and the most rostral projection of the third ethmoturbinate. Olfactory neuron loss in the 30 ppm (42 mg/m³) exposure animals was mild (grade 1) to moderate (grade 2), but olfactory neuron loss in the 80 ppm (111 mg/m³) exposure group increased to an average severity between moderate (grade 3) and severe (grade 4). Basal cell hyperplasia was observed in both exposure groups at this level of the nasal cavity but was slightly more severe in the 30 ppm (42 mg/m³) exposure group.

Olfactory mucosa lines most of the nasal cavity at levels 5 and 6 while the ventral meatus and lateral walls of the nasal cavity are lined with respiratory epithelium. Although the olfactory mucosa was widely distributed, lesions in this tissue were found at select sites. At level 5, mild

to moderate olfactory neuron loss and mild basal cell hyperplasia mainly affecting the nasal septum, dorsal nasal cavity, and marginal ethmoturbinate were observed in both exposure groups. The nasal septum was not affected in the 30 ppm (42 mg/m³) exposure group. The same pattern and severity of lesions were observed at level 6 in the 80 ppm (111 mg/m³) exposure group.

The critical effects in this study are nasal lesions of the olfactory mucosa, 30 ppm (42 mg/m³) is the LOAEL and 10 ppm (14 mg/m³) the NOAEL (Brenneman et al., 2000).

It should be noted that lesions in the nasal tract reported in other studies are not completely in agreement with those reported by Brenneman et al. (2000) where lesions were clearly related to dose and localized in the olfactory epithelium. In mice exposed to similar levels of H_2S (CIIT, 1983a), inflammation of the nasal mucosa was reported and described as occurring primarily in the squamous nasal mucosa and extending into the respiratory epithelium, with no specific mention of olfactory mucosal involvement. Sectioning of the nasal epithelium was through 4- rather than 6-areas examined by Brenneman et al. (2000) explaining, in part, this difference. Effects in the nasal cavity reported in Sprague-Dawley rats in an earlier study (CIIT, 1983c) differ in severity and location from those reported by Brenneman et al. (2000). Reasons for such a discrepancy are unknown.

4.3. REPRODUCTIVE/DEVELOPMENTAL STUDIES—ORAL AND INHALATION

Dorman et al. (2000) attempted to determine if perinatal exposure by inhalation to H_2S had an adverse effect on reproductive performance and pregnancy outcomes. Virgin male and female Sprague-Dawley rats (12/sex/group) were exposed to 0, 10, 30, or 80 ppm (0, 14, 42, or 111 mg/m³) H_2S 6 hr/day, 7 days/week for two weeks prior to breeding. Exposure continued during a 2-week mating period and throughout gestational days 0-19 (GD 0-19). Evidence of copulation (vaginal plug or sperm in vaginal lavage fluid) during the 2-week mating period was considered GD 0. On postnatal day (PND) 4, litters were randomly reduced to 4 animals per sex when possible. Remaining pups were euthanized and discarded without being examined. Dams and pups were then exposed PND 5-18. Nonpregnant adult females were exposed for an additional 23-24 days following the 2-week breeding period. Adult males were exposed to H_2S for 70 consecutive days.

Clinical examinations were performed on all animals before and after each exposure (Dorman et al., 2000). The body weights of the F_0 males and females were determined weekly throughout the study, except that female body weights were not determined weekly once evidence of mating was present. Presumed pregnant females were weighed on GD 0, 7, 14, and 20, and dams were weighed on PND 0, 4, 7, 14, and 21. Feed consumption was determined weekly in F_0 males and prebreeding females. Feed consumption in presumed pregnant females

was recorded for GD 0-7, 7-14, and 14-20. Dam feed consumption was recorded for PND 0-4, 4-7, 7-14, and 14-21. At the end of exposure, adult rats were euthanized and a complete necropsy was performed with emphasis on reproductive and accessory sex organs. Post-parturient animals were necropsied the day of or the day after weaning. At necropsy, the right testis from each F_0 male was examined for sperm number, production, motility, and morphology.

No deaths or adverse clinical signs were observed in F_0 males and females for any exposure group. There was a statistically significant decrease in feed consumption in male rats exposed to 80 ppm (111 mg/m³) H₂S during the first week of the study. There was a small, but not statistically significant, decrease in body weight (5-6%) observed in F_0 males and females exposed to 80 ppm (111 mg/m³) H₂S that was present throughout entire exposure period. The only significant difference in organ weights was an increase in absolute and relative adrenal gland weights observed in F_0 males exposed to 10 ppm (14 mg/m³) and 80 ppm (111 mg/m³) but not in the mid-dose of H₂S and a decrease in the relative ovary weight of females in the 10 ppm (14 mg/m³) exposure group.

There were no statistically significant effects on reproductive performance (mating index, fertility index, postimplantation loss per litter, and number of late resorptions or stillbirths) in F₀ animals. Also, the number of live pups, litter size, average length of gestation, and average number of implants per female were not affected. There were no statistically significant effects on sperm production or morphology noted in F₀ males; however, a large percentage of abnormal sperm was observed in one F_0 male from both the 30 (42 mg/m³) and 80 ppm (111 mg/m³) exposure groups (29 and 73%, respectively). In comparing high-exposure and control animals, there were a few histologic diagnoses with a higher incidence in the high-exposure group, including: testicular tubular degeneration, intratubular sperm stasis, tubular mineralization, sperm granulomas, and multinucleated giant cells, degenerate sperm forms in the lumen, aspermia, and oligospermia. While the incidence of testicular tubular degeneration in highexposures was higher (42%) than the controls (17%), statistical significance was not achieved. One incidence each of sperm granuloma and unilateral necrosis of the cauda was present in the 80 ppm (111 mg/m³) exposure group. Notable histological findings in females included one each of ovarian cysts and squamous metaplasia of the endometrium in the 10 ppm (14 mg/m^3) and 30 ppm (42 mg/m^3) exposure groups. The olfactory neuron loss and basal cell hyperplasia observed in the male rats in this study are reported elsewhere (Brenneman et al., 2000). The assignment of any effect level for reproductive toxicity is problematic since the study considered only 12 rats/sex/group and the study also lacked any statistical significance in either function or histopathology related to reproduction endpoints. In summary, the reproductive segment of this study did not demonstrate reproductive toxicity in either male or female rats following relatively high H_2S exposure of 80 ppm (111 mg/m³) under repeated conditions (up to 70 days for males).

With regards to developmental effects from H_2S exposure, there were no statistically significant increases in structural malformations. Observed malformations included kinked tail, agenesis of the tail, anophthalmia, small rear legs and body, frontal bone defects, hypognathia, and skin lesion characterized by detachment of the skin and dermis. However, none of these effects was dose-related. There were no significant differences in pup weight gain or development (pinnae detachment, surface righting, incisor eruption and negative geotaxis, vaginal patency, preputial separation, and eyelid separation). Surface righting was also equivalent across exposure groups. There were no treatment-related effects on motor activity, acoustic startle response, passive avoidance observed, or FOB. Terminal body and organ weights in all exposure groups were comparable to controls. Although a wide variety of gross observations were noted, they were not considered treatment-related by the investigators. Microscopic examination of six levels of the brain from pups from the control and high-dose group failed to demonstrate any histologic abnormalities.

Saillenfait et al. (1989) reported that the exposure of pregnant Sprague-Dawley rats to 0, 70, 140 or 210 mg/m³ (0, 50, 100, or 150 ppm) H₂S 6 hr/day during gestational days 6-20 resulted in no signs of maternal toxicity or adverse effects on the developing fetus. While there was a slight but significant decrease in fetal weight in all treated groups, the biological significance of this finding is questionable since the dams also lost weight and there were greater numbers of live fetuses per litter in those groups.

Significant elevations in maternal blood glucose levels were observed in Sprague-Dawley rats exposed to 0, 28, 70, or 105 mg/m³ (0, 20, 50, or 75 ppm) H₂S, 7 hr/day from gestation day 1 through postnatal day 21 (Hayden et al., 1990a). There were no changes in serum protein, LDH, SGOT, or alkaline phosphatase activities in the treated dams or pups. Pups exhibited decreased serum triglycerides in the 70 mg/m³ (50 ppm) dose group. In a similarly-designed study using the same exposure levels (Hayden et al., 1990b), there was a dose-dependent increase in parturition time of approximately 10, 20, and 40% over matched controls at 0, 28, 70, or 105 mg/m³ (0, 20, 50, or 75 ppm), respectively. The biological significance of this latter effect is questionable since parturition time was variable among control groups and was not examined statistically. Maternal liver cholesterol content was significantly elevated on day 21 postpartum in the high-dose dams.

Although no studies have reported effects that prenatal exposure to H_2S may eventually have upon the adult nervous system, early postnatal measurements provide indications of possible neurotoxic effects that may occur in adulthood. Hannah et al. (1989) exposed timedpregnant Sprague-Dawley rats (number not reported) to 105 mg/m³ (75 ppm) H_2S , 7 hr/day from day 5 postcoitus until PND day 21. Groups of 8 randomly-selected offspring, exposed *in utero* through PND 21, were sacrificed on PNDs 7, 14, and 21. The levels of putative neurotransmitter aspartate and gamma amino-butyric acid (GABA) were all significantly (p<0.05) depressed in the cerebellum on PND 21. Glycine was unaltered. GABA is an inhibitory neurotransmitter of the central nervous system that appears in Purkinje cells such that decreases in GABA could be due to a decrease in the population of Purkinje cells. Analysis of cerebellum (in the 75 ppm animals only), however, established that this was not the case. There were no significant differences in the brain weights of treated pups compared to controls. The significance of these amino acid changes to neurobehavioral development is uncertain.

In a similarly designed study using the same exposure protocol, Hannah and colleagues (Skrajny et al., 1992) exposed groups of 20 pregnant Sprague-Dawley rats to 0, 28, or 105 mg/m³ (0, 20, or 75 ppm) H₂S. Ten randomly-selected pups, exposed *in utero* through PND 21, were sacrificed on PND 7, 14, and 21. Significantly increased (p<0.05) serotonin levels were observed in the frontal cortex on postpartum day 21 in pups exposed to 28 mg/m³ (20 ppm) H₂S, while increased (p<0.01) serotonin levels were observed in the cerebellum and frontal cortex on postpartum days 14 and 21 in pups exposed to 105 mg/m³ (75 ppm) H₂S. Norepinephrine levels increased (p<0.05) at 105 mg/m³ (75 ppm) in the cerebellum on PNDs 7, 14, and 21, and in the frontal cortex on PND 21. At 28 mg/m³ (20 ppm), frontal cortex norepinephrine levels were decreased compared to controls on days 14 and 21.

In a study designed to evaluate the perinatal effect of H₂S on developing cerebellar Purkinje cells, Sprague-Dawley dams were exposed to 28 or 70 mg/m³ H₂S (20 or 50 ppm) for 7 hr/day from postcoital day 5 until PND 21 (Hannah and Roth, 1991). Control animals were exposed similarly, but without H₂S. Pups born to treated dams were also exposed perinatally up to PND 21 (Roth, 2002), at which time one pup was randomly selected from each litter for analysis, resulting in a total of 10 pups for each concentration and 10 pups from each of the control groups. Cerebella were processed using the Golgi method. One Purkinje cell from the pyramis of each cerebellum was digitized for morphometric analysis. Quantitative analysis of dendritic morphology is generally regarded as a traditional morphometric method appropriate to assess the effect of environmental influences, including chemical insult, on the developing brain. The authors used a vertex analysis (after Berry and Flinn, 1984) that can be used on a small number of nodal vertex types. This method leads to a determination of random versus nonrandom growth. Statistical evaluation was carried out using the Student's *t*-test. Exposure to both 20 and 50 ppm (28 or 70 mg/m³) produced significantly longer segment length than controls over the low and middle branching orders, but not the higher orders. It was stated that segment lengths are typically increased when afferent fibers are experimentally reduced in number. The mean vertex path length (defined as a measure of the mean distance from any one equivalent order back to the axon hillock) was found to increase (p<0.001) in incremental steps in both treated groups compared to controls. When considered over the entire dendritic tree, these changes could alter the capacity of Purkinje cells to integrate afferent impulses and may be associated with an increased likelihood of an abnormal physiological response. Segment

branching was also found to be different in cells from the treated animals. There was a significant increase in branch number of the middle orders and a significant decrease in numbers over the higher orders (considered to represent the 'growth front').

The investigators considered this to be a compensatory response when there is restriction on the 'growth front'. The question of random versus nonrandom growth patterns was evaluated by examining the symmetry of the branching. It was found that vertex analysis indicated that the treated cells were undergoing a significant amount of nonsymmetrical or nonrandom growth, suggesting that H_2S affected the branching pattern. The resulting alterations in dendritic arborization appeared to be in response to H_2S exposure at the 20 and 50 ppm (28 or 70 mg/m³) levels such that the 20 ppm (28 mg/m³) level may be considered a low-effect rather than a noeffect level.

The significance of the findings in this and the Skrajny et al. (1992) studies to humans is unclear. Since alterations may constitute a chemically-induced change in the growth or organization of structural (or neurochemical) elements, they are to be regarded as possible indicators of a neurotoxic effect in accordance with guidance in EPA's neurotoxicity risk assessment guidelines (U.S. EPA, 1998). The question as to whether the alterations are adverse, however, remains unclear for several reasons. The effects reported by Hannah and Roth (1991), for example, are highly selective and could be due to environmental factors not directly related to exposure including variability resulting from the restricted sampling technique (i.e., one Purkinje cell/pup). Also, the relationship of these alterations to a functional impairment is unclear and cerebellar neurological sequelae are not the most obvious symptom associated with H_2S exposure. It is acknowledged, however, that predicting functional impairments from specific structural alterations such as reported by Hannah and Roth (1991) is difficult due to the selective nature of the observed alterations and the dynamic self-organizing response of the developing brain to injury.

The effect of H_2S on spermatogenesis in the rat was assessed by Andrew et al. (1980). A group of 10 14-week-old male Wistar rats were exposed to 220 ppm (306 mg/m³) H_2S for 3 hr/day for 7 days, compared to a control or nonexposed group. Following the last exposure, each male was caged with two unexposed virgin females for 10 consecutive weeks. Four male rats were administered triethylenemelamine and used as positive controls. The females were sacrificed after 18 days cohabitation. Parameters evaluated included fertility, corpora lutea, total implants, and dead implants. There were no adverse effects of exposure on these parameters. Subsequently, the investigators evaluated exposure on prenatal development by exposing pregnant Wistar rats to 220 ppm (306 mg/m³) for 3 hr/day, for 5 days either days 1-18 of gestation, or days 7-11 or 12-16 (5 days total). All animals were sacrificed at day 21 and examined for dead implants, fetal malformations, and growth retardation. There was no evidence of embryotoxicity or external malformations.

4.4. OTHER STUDIES

4.4.1. Respiratory Toxicity

Cytotoxic effects of H₂S exposure in the respiratory tract of rats was investigated under acute exposure conditions in a series of studies by Lopez et al. (1987). Fischer 344 rats were exposed to 0, 10, 200, or 400 ppm (0, 14, 278, or 556 mg/m³) H₂S for 4 hours and then sacrificed at 1, 20, or 44 hours postexposure (Lopez et al., 1987). Bronchoalveolar lavage (BAL) and nasal lavage (NL) were performed on all animals. Lavage protein, lactate dehydrogenase (LDH), alkaline phosphatase (AP), cellular composition, y-glutamyl transpeptidase (GGT), and cytopathology were determined. Animals exposed to 400 ppm (556 mg/m³) H_2S exhibited a significant, but transient increase (800%) in NL protein at 1 hour postexposure that returned to normal at 20 hours postexposure. There was also a transient increase (320%) in LDH from NL fluid in the 400 ppm (556 mg/m³) exposure group. The cellularity in the NL fluid was increased by 139 and 483% in the 10 and 200 ppm (14 and 278 mg/m³) 1 hour postexposure groups, but these values returned to baseline by 20 hr postexposure. Cellularity was increased by 817, 501, and 73% at 1, 20, and 44 hours postexposure, respectively, for animals exposed to 400 ppm (556 mg/m^3) H₂S. Differential cell counts indicated that the increase in cellularity was initially due to desquamation of nasal epithelial cells (1 hour postexposure) followed by exudation of polymorphonuclear lymphocytes (PMN). Cytopathology revealed that the desquamated epithelial cells showed stages of ciliocytophthoria, cytoplasmic vacuolation, pyknosis, karyorrhexis, cytoplasmic constriction between the basal region and terminal plate, and separation of ciliated tufts. In the 10 ppm (14 mg/m³) and 200 ppm (278 mg/m³) animals, no other effects in NL fluid were observed.

Similar to NL fluid, the BAL fluid of rats exposed to 400 ppm (556 mg/m³) had an increase in protein content of approximately 3000, 1400, and 450% at 1, 20, and 44 hours postexposure. There was no increase in BAL fluid proteins in the other exposure groups. In animals exposed to 400 ppm (556 mg/m³) H₂S, LDH significantly increased at all postexposure times, but LDH decreased toward baseline at the 20 and 44 hour postexposure times. In the 200 ppm (278 mg/m³) exposure group, LDH activity in the BAL fluid was significantly decreased in the 44 hr postexposure group. In animals exposed to 400 ppm (556 mg/m³) H₂S, AP initially significantly increased in the 1 hour postexposure group, but significantly decreased below baseline at 20 and 44 hours postexposure. Animals exposed to 200 ppm (278 mg/m³) H₂S had AP levels below baseline at 20 and 44 hours postexposure. GGT was elevated (900%) in the 400 ppm (556 mg/m³) 1 hour postexposure animals. Animals exposed to 400 ppm (556 mg/m³) H₂S had a 42% reduction in nucleated cells in the BAL fluid at 1 hour postexposure, but which returned to normal at 20 hours and increased to 30% over baseline at 44 hours postexposure.

Pulmonary alveolar macrophages were significantly reduced at 1 hour postexposure in animals exposed to 400 ppm (556 mg/m³) H₂S, but returned to normal by 20 hours postexposure. PMN increased in all postexposure times (highest at 20 hours) in animals exposed to 400 ppm (556 mg/m³) H₂S. In the 10 ppm (14 mg/m³) and 200 ppm (278 mg/m³) animals, no other effects in BAL were observed.

The investigators concluded that vacuolization, ciliocytophthoria (a cellular degenerative change often from accompanying viral infections), and nasal sloughing is evidence of cytotoxicity of the nasal epithelium. The investigators also cautioned that although cell count in the NL fluid was elevated in animals exposed to 10 ppm (14 mg/m³) H_2S , dramatic changes can rapidly be restored due to the remarkable repair capacity of the respiratory epithelium. In addition, the cell counts in rats exposed to 10 ppm (14 mg/m^3) returned to baseline values by postexposure hour 20. Exudation of protein into the nasal passages is a sign of vascular permeability, and the investigators suggested that increased LDH in the NL fluid is a sign of damage to the epithelium. The investigators also suggested that increased LDH and AP observed 1 hour postexposure in BAL fluid is suggestive of damage to the pulmonary epithelium. However, the investigators stressed that H₂S appears to be a weaker toxicant to the respiratory epithelium than other pneumotoxicants. The authors stated that the olfactory epithelium appears to be more sensitive to the toxic effects of H₂S than the respiratory epithelium. In addition, the investigators considered the large increase in protein content in the BAL to be consistent with the observed edematous properties of the gas. The LOAEL in this study is 200 ppm (278 mg/m³). The NOAEL is 10 ppm (14 mg/m³).

Histologic and ultrastructural alterations in the lungs of rats were reported in a similar study in which male Fischer 344 rats were exposed for 4 hours to either 82 or 440 ppm (116 or 613 mg/m^3 ; n = 12 rats per dose level) H₂S followed by sacrifice at 1, 18, or 42 hours postexposure (Lopez et al., 1988a). Histologic changes were transient and mainly present in rats exposed to 440 ppm (613 mg/m³) H_2S . While some histologic changes were noted at 82 ppm (116 mg/m^3) , no pathologic changes were reported at this exposure level. In rats exposed to 440 ppm (613 mg/m³) H_2S , bronchiolar ciliated cells developed necrosis, but necrotic damage was rapidly repaired through mitosis. Examination of the lung demonstrated the notable edematogenic effect following exposure to 440 ppm (613 mg/m³) H₂S. Based on findings from the postexposure groups, the investigators suggested a chronology to the edematogenic effect in which fluid first accumulates around the blood vessels, then the interstitium, and finally in the alveoli. Perivascular edema without involvement of the alveoli in animals exposed to 82 ppm $(116 \text{ mg/m}^3) \text{ H}_2\text{S}$ suggested that fluid that accumulated in the interstitium was reabsorbed before it entered the bronchoalveolar spaces. This may explain the finding by Lopez et al. (1987) that protein was detected in BAL fluid from animals exposed to 400 ppm (556 mg/m³) H_2S , but not 200 ppm (278 mg/m³) H_2S . The investigators also found a lack of structural changes in the

alveolar endothelium, basement membrane, or type I pneumocytes, suggesting that H_2S exposure to concentrations as high as 440 ppm (613 mg/m³) does not compromise the air-blood barrier. In addition, there was no adverse effect on the mast cells. If damage to the air-blood barrier and mast cell degranulation is not responsible for the observed pulmonary edema following H_2S exposure, then the investigators suggested that the edema may be due to an outflow of liquid from the peribronchovascular connective tissue into the lumen of small airways via high conductance pathways that fill the alveolar spaces in a retrograde manner.

Lopez et al. (1988b) examined the effect of acute H_2S exposure on nasal epithelial cells. Male Fischer 344 rats were exposed to 0, 14, 280, or 560 mg/m³ (0, 10, 200, or 400 ppm) H_2S for 4 hours. Animals were sacrificed at 1, 18, or 44 hours postexposure. No animals died, but clinical signs of lethargy and epiphora were present in animals exposed to 560 mg/m³ H_2S . Nasal lesions were present only in the high-dose group and manifested as necrosis and exfoliation of the respiratory and olfactory epithelium. Of the four different sections of the nasal cavity, sections two and three (mid-nasal cavity) were the most severely affected. The rostral section (section 1) was not affected, and section 4 was only slightly affected. These results are similar to those reported by Brenneman et al. (2000). At 44 hours postexposure, the respiratory mucosa was essentially repaired, but the olfactory mucosa continued to exfoliate. Effects of H_2S on nasal parameters were much less severe in the Sprague-Dawley rat study reported by CIIT (1983c).

Green et al. (1991) examined the effects of H_2S from the BAL of rats under the hypothesis that the hemorrhagic pulmonary edema observed from exposure to high levels of H_2S is associated with changes in pulmonary surfactant. Male Fischer 344 rats were exposed to 0, 200, or 300 ppm (0, 278, or 417 mg/m³) H_2S for 4 hours, and at 1 hour postexposure BAL fluid was obtained. Marked abnormalities in surfactant activity were demonstrated in the lavagates from rats exposed to the highest concentration only 300 ppm (417 mg/m³) and not in the lavagate from the lower dose 200 ppm (278 mg/m³) or the controls. The lungs of high-dose animals showed areas of red atelectasis, patchy alveolar edema, and perivascular edema. These results are suggestive of a threshold for this surfactant response and subsequent histopathological progression.

The effects of H_2S on lung bacterial defense has been investigated. Rogers and Ferin (1981) exposed male Long-Evans rats to 45 ppm (636 mg/m³) H_2S for 2, 4, or 6 hr followed by bacterial challenge to *Staphylococcus epidermidis*. In control animals, most of the bacteria was inactivated by the 6-hour postchallenge sacrifice time. Rats exposed to H_2S for 2 hours responded similarly to controls. However, rats exposed to H_2S for 4 and 6 hours had 6.5- and 52-fold greater percent bacteria remaining, respectively, compared to controls. The investigators suggest that an H_2S -induced absence of bacterial inactivation may explain secondary pneumonias in humans subsequent to acute or subacute H_2S exposure. The effect of bacterial

inactivation was hypothesized by the investigators to be due to alveolar macrophage inactivation. The hypothesis is supported by Robinson (1982) who demonstrated that rabbit alveolar macrophages lost the phagocytic ability *in vitro* when exposed to 54 ppm (75 mg/m³) H_2S for 24 hours.

Although the above studies are acute in nature (i.e., mostly 4 hr) and are not appropriate for the establishment of a chronic guideline value, the results may offer insights into the processes underlying affects observed in the respiratory tract tissues in animals exposed in a more protracted manner to H_2S such as by Brenneman et al. (2002).

Nasal tissues were examined for histology and cytochrome oxidase immunoreactivity in 10-week-old male CD rats 24 hours after exposure to 0, 30, 80, 200, or 400 ppm (0, 42, 112, 280, or 560 mg/m³) H₂S for 3 hr/day for 1 day or for 5 consecutive days (Brenneman et al., 2002). The nose was histologically examined 24 hours after exposure, and lesion recovery was assessed at 2 and 6 weeks following the 5-day exposure. The single 3-hour exposure to ≥ 80 ppm H₂S resulted in regeneration of the respiratory mucosa and full thickness necrosis of the olfactory mucosa localized to the ventral and dorsal meatus, respectively. Repeated exposure to the same concentrations caused necrosis of the olfactory mucosa with early mucosal regeneration that extended from the dorsal medial meatus to the caudal regions of the ethmoid recess. Acute exposure to 400 ppm H₂S induced severe mitochondrial swelling in sustentacular cells and olfactory neurons, which progressed to olfactory epithelial necrosis and sloughing. Cytochrome oxidase immunoreactive cells were more frequently observed in regions of the olfactory mucosa commonly affected by H₂S than in regions that were not. These findings demonstrate that acute exposure to >80 ppm H₂S resulted in reversible lesions in the respiratory and olfactory mucosae of the CD rat and, further, that cytochrome oxidase activity (as judged from cytochrome oxidase immunoreactivity) may be a susceptibility factor for H₂S-induced olfactory toxicity in the CD rat.

Transmission electron microscopy of cultured rabbit alveolar macrophage (AM) cells exposed *in vitro* to 54 ppm (75 mg/m³) for 24 hours revealed extensive damage to interior organelles and membrane resulting in a decreased percentage of AM capable of phagocytosis (Richardson, 1995). Khan et al. (1991) found evidence that H₂S exposure of F344 rats decreased *in vitro* respiratory rates of collected AM cells (zymosan-stimulated) and at the highest concentration decreased cell viability. Rats had been exposed to 0, 70, 280, or 560 mg/m³ (0, 50, 200 or 400 ppm) for 4 hours before cell collection.

Moulin et al. (2002) correlated nasal pathology in H_2S -exposed rats with regions of high flux in the rat nasal cavity generated from anatomically accurate computational fluid dynamics (CFD). The specific location of lesions from nasal tract sections approximating sections 4 and 5 of a study of Brenneman et al. (2000) were noted and the predominately olfactory-lined portions of the perimeter of these 2 levels were divided into roughly similar size regions. CFD flux of

 $\rm H_2S$ at these regions and the incidence of nasal pathologies within these regions were then calculated and correlated. The analysis showed that $\rm H_2S$ -related lesions were observed only at specific sites within the olfactory epithelium and that areas of respiratory epithelium were not affected. Further, the incidence, severity and distribution of lesions seen within the olfactory epithelium increased with concentration (approximately 50% of the olfactory epithelium affected at 30 ppm [42 mg/m³] and 70% at 80 ppm [111 mg/m³]) whereas the olfactory epithelium of rats exposed to 10 ppm (14 mg/m³) and controls were unaffected. In contrast to the localization of nasal lesions in the olfactory epithelium, CFD simulations predicted higher flux (and therefore a higher dose) to those regions lined by respiratory epithelium where no H₂S-related lesions were observed. Region by region ranking of lesion incidence in the olfactory-lined epithelium with predicted flux revealed a high degree of correlation (Spearman's rank correlation coefficient, p < 0.005 for level 5 section) for the 30 and 80 ppm exposure groups.

Results from Moulin et al. (2002) suggest that olfactory epithelium is specifically sensitive to the effects of H_2S while respiratory epithelium is resistant. These results therefore lend credence to considering nasal toxicity from H_2S as relevant to humans having olfactory epithelium that can be impinged upon by inhaled air streams containing H_2S . The high correlation between predicted H_2S flux (a surrogate of dose) in the olfactory epithelium and nasal response (histopathology) indicate dose-response analysis should consider such information. Such procedures could be used, for example, to extrapolate dose-response relationships between animal models and humans to accommodate the differences between these species, i.e., the relatively large area of olfactory epithelium in areas of high flux in rats and mice (Kimbell et al., 1997) compared to the relatively small area of olfactory epithelium in areas of low flux in humans (Subramaniam et al., 1998). In considering this possibility, however, the authors caution that more refined estimates of necessary parameters such as nasal extraction efficiency would be required before such a procedure could be applied in a quantitative manner.

4.4.2. Neurotoxicity

Kombian et al. (1988) examined the acute effects of H_2S on amino acid levels in various brain regions of the rat following i.p. administration of 10 or 30 mg/kg doses. Hydrogen sulfide administration was associated with alterations in several amino acid transmitters, in particular, enhanced levels of aspartate, glutamate, glutamine, GABA, glycine, taurine and alanine in the brainstem. Amino acid transmitters may play a role in the neuronal control of breathing. Kombian et al. (1988) postulated that the observed changes in neurotransmitter amino acid levels in the brainstem may at least partially contribute to the acute toxicity of H_2S through mechanisms involving respiratory arrest. Hannah et al. (1989; see Section 4.3) also reported depressed levels of these same neurotransmitters (save for glycine) in the cerebella of 21-day-old rats that had been exposed *in utero* and up to postnatal day 21 to airborne levels of H_2S at 105 mg/m³ for 7 hr/day. Elovaara et al. (1978) demonstrated a marked reduction in the incorporation of labeled leucine in cerebral protein and myelin in adult female mice exposed to 100 ppm (139 mg/m³) H_2S for 2 hours. The results indicate decreased brain protein synthesis. The authors suggested that the changes may have resulted from the inhibition of cerebral CytOx activity. The authors also noted that 100 ppm (139 mg/m³) is 5 to 20 times lower than exposure levels associated with human fatalities.

Utilizing *in vitro* whole brain preparations, Roth et al. (1997) showed that the addition of 0.13μ M Na₂S can inhibit CytOx and carbonic anhydrase. Also, KCN, a specific inhibitor of cytochrome oxidase, altered the rhythmical activity of rat hippocampal neurons in a manner similar to Na₂S or H₂S. The authors speculated that the inhibition of the respiratory enzyme CytOx may be involved in the sulfide-induced alteration of neuronal function in brain regions such as the hippocampus.

Skrajny et al. (1996) investigated the effects of low H₂S levels on the hippocampus and neocortex of freely moving rats using electroencephalographic activity (EEG) as a measurement of neuronal function. The hippocampus was a focus of this study because this exposure of humans to high levels of H₂S (i.e., approximately 500 ppm or 695 mg/m³) has been associated with memory loss and varying degrees of learning and perceptual deficits (Kilburn, 1993; Tvedt et al., 1991a,b; Wasch et al., 1989). The hippocampus plays a role in processing between cortical structures involved in cognitive behavior, and is very susceptible to toxic insult. Rats were exposed to 25, 50, 75 or 100 ppm (35, 69, 104, or 139 mg/m³) H_2S for 3 hr/day for 5 days. Results, recorded during the last 10 minutes of the exposures, indicated that repeated exposure to H₂S produced a dose-related increase in the amplitude of hippocampal theta activity the total of which was cumulative with exposure duration. The effect of H₂S on hippocampal EEG was dose-dependent and highly significant at all concentrations. The authors postulated that elevated catecholamine and/or serotonin levels from the inhibition of MAO may be involved in the sulfide-induced alteration of hippocampal theta activity, since the time for reactivation of MAO activity correlated with the time required for the recovery of the total power of hippocampal EEG activity following exposure to H_2S . The LOAEL for this study is 35 mg/m³ (25 ppm). The authors also noted that exposure to 100 ppm (139 mg/m³) H_2S produced no significant effects on rat EEG activity recorded from the neocortex.

In an earlier study, Skrajny et al. (1992) examined the effect of H_2S exposure on serotonin and norepinephrine levels in the developing cerebellum and frontal cortex of Sprague-Dawley rats. Timed-pregnant rats were exposed to 20 or 75 ppm (28 or 104 mg/m³) for 7 hr/day from day 5 postcoitus until PND 21. At parturition, treated and control animals were culled to 6 dams, with a maximum of 12 pups per litter with 10 pups in each group randomly selected for euthanization at PND 7, 14, and 21. The number of litters in each group of dams was not stated nor was there an indication if pups were selected from randomly selected litters. Exposure to 75 ppm (104 mg/m³) resulted in a statistically significant (p<0.01) increase in serotonin in the cerebellum from PND 7 to 14, compared to controls. At PND 21, the value remained unchanged from PND 14. The increase was even more pronounced in the frontal cortex from PND 7 through 21 (p<0.01). The increase at 20 ppm (28 mg/m³) in the frontal cortex was less pronounced than at 75 ppm (104 mg/m³) and reached statistical significance only at PND 21 (p<0.05). There was no significant increase in the level in the cerebellum. Changes in norepinephrine were more difficult to interpret as there was no clear pattern of change. The investigators concluded that the effects of exposure on serotonin and norepinephrine, coupled with the laboratory's prior evidence demonstrating possible morphological changes in Purkinje cell development and the known influence of monoamines on neural development, suggest a mode of action for H₂S on the CNS.

Higuchi and Fukamachi (1977) examined the effects of inhalation exposure to H_2S on avoidance behavior in rats. Rats were exposed to 100–500 ppm (139-695 mg/m³) H_2S . Hydrogen sulfide was generated from the reaction of iron sulfide and hydrochloric acid. The exposure duration was 2 hours. Both the Sidman-type conditioned avoidance and discriminated avoidance behavior were assessed. A rapid and significant decrease in discriminated avoidance response was observed at concentrations ≥ 200 ppm (278 mg/m³). At concentrations ≥ 300 ppm (417 mg/m³), the Sidman-type conditioned avoidance response was also decreased. The behavioral effects were reversed following ventilation with clean air or when exposure was terminated.

In an attempt to demonstrate that the neurological sequelae following exposure to high concentrations of H₂S is due to neuronal necrosis, Baldelli et al. (1993) administered sodium sulfide (Na₂S) to unventilated and anesthetized-ventilated (to compensate for the loss of central respiratory drive) male Wistar rats by i.p. injection. Doses in unventilated rats ranged from 84 to 200 mg/kg, while ventilated rats were administered 120, 150, or 200 mg/kg. In unventilated animals, all animals that were administered doses greater than 120 mg/kg (n=11) sulfide died in \leq 10 minutes. Doses of 120, 108, 100, 96, and 84 mg/kg were lethal to 7/10, 3/3, $\frac{1}{2}$, 2/5, and 0/1 animals, respectively. An LD₅₀ of 94 mg/kg was calculated for 32 unventilated rats by the investigators. Of the 8 surviving unventilated animals, only one 120 mg/kg rat demonstrated histopathological signs of neuronal necrosis of the cerebral cortex. In ventilated animals (5 rats/group), an LD₅₀ of 190 mg/kg was calculated. None of the 120 or 150 mg/kg group animals died during compound administration. However, 4/5 of the 200 mg/kg group animals died within minutes of compound administration, and the sole-surviving 200 mg/kg animal and 4/5 of the 150 mg/kg group animals died within one week of exposure. Blood pressure and arterial blood gases were monitored only in ventilated animals. None of the ventilated animals exhibited a decrease in arterial pO₂ or alteration in pH or pCO₂. Mean arterial blood pressure decreased to 42.8, 34.2, and 16.8 mm Hg for ventilated animals within 2-5 minutes of compound

administration. Mean arterial blood pressure recovered within 15- 20 minutes for the 120 and 150 mg/kg dose groups, but blood pressure did not return to baseline for 50 minutes in the lone surviving 200 mg/kg ventilated animal. There were significant deceases in EEG activities that were also dose-dependent. After doses of \geq 150 mg/kg, EEG activity decreased to near isoelectric levels. EEG activity returned to normal in the two lowest doses and showed some recovery at the highest dose. Neuronal necrosis of the cerebral cortex, caudate, and brain stem was observed in the surviving 200 mg/kg animals. The investigators concluded that neuronal necrosis is not directly related to H₂S exposure. Rather, neuronal necrosis is indirectly caused by hypotension induced by H₂S exposure. The profound hypotension observed in the surviving 200 mg/kg ventilated in cerebral ischemia.

Three reports describe effects of H₂S on neurological function in rats (Partlo et al., 2001; Struve et al., 2001). Partlo et al. (2001) exposed adult rats to125 ppm (174 mg/m³) H₂S for 4 hr/day for 5 days/week for 5 weeks. Controls were exposed to a nitrogen/air mixture. Overt signs of eve irritation, respiratory distress, behavioral dysfunction or impaired consciousness were not observed at any time during the experiment. Testing of the animals revealed that this level of exposure did not affect memory retention or acquisition, but did impair one measure of performance during the reacquisition of a reversed contingency (radial arm maze task). These data provide evidence of mild brain dysfunction following a prolonged period of repeated hydrogen sulfide at this exposure level. The authors discuss the prefrontal cortex as a potential target site. Struve et al. (2001) exposed adult rats to the inhalation of 0, 30, 80, 200, or 400 ppm $(0, 43, 111, 278, \text{ or } 556 \text{ mg/m}^3)$ H₂S for 3 hr/day for 5 consecutive days. When exposure levels equaled or exceeded 80 ppm (111 mg/m³), the animals exhibited significant reductions in motor activity, body temperature, and water maze performance (spatial learning); it is to be noted, however, that the latter result may have been confounded by the former. Pregnant Sprague-Dawley rats (12/group) were exposed to either 0, 10, 30, or 80 ppm $(0, 14, 42, \text{ or } 111 \text{ mg/m}^3)$ H₂S throughout gestation and postnatally to both dams and pups (8 rats/litter after culling) PND 5 - 18. No decrement in behavioral performance, as judged by testing for motor activity (conducted on PNDs 13, 17, 21, and 60), passive avoidance (conducted on PND 22 and 62), functional observational battery (conducted blind on PND 60), and acoustic startle response (conducted on PND 21 and 62), was noted in the offspring at any exposure concentration (Dorman et al., 2000).

4.4.3. Cardiovascular Toxicity

In a series of experiments, Kosmider et al. (1967) exposed 10 rabbits to 100 mg/m³ (72 ppm) H_2S for 1.5 hours and 17 rabbits to 100 mg/m³ for 0.5 hours for 5 days. Electrocardiograms were taken after exposure and histochemical studies of the heart were performed. Ten rabbits served as controls for the histochemical studies and control electrocardiograms were made on all animals 10 days prior to exposure. Nine of ten animals exposed for a single 1.5 hour exposure exhibited mainly disorders in ventrical repolarization (T wave depression or inversion) without cardiac arrhythmia. However, animals exposed for 5 days exhibited arrhythmias such as ventricular extrasystoles and bigeminal rhythm or contractions elicited from several pacemakers. These effects were noted in 15 of 17 rabbits, and two of the affected animals exhibited atrial fibrillation. The arrhythmias lasted for several days after exposure. In addition, multiply exposed animals also exhibited T wave depression. Histochemical studies revealed a decrease in ATP phosphohydrolase and NADPH₂ oxidoreductase activity, which suggests a direct effect on myocardial cells. A direct effect on ATP phosphohydrolase could alter Na⁺ and K⁺ transport to the myocardial cell and could result in arrhythmia. Administration of sodium citrate, a calcium chelator, attenuated the arrhythmias.

Electrocardiogram effects have also been reported in rats. Kohno et al. (1991) exposed male Wistar rats to 75 ppm (105 mg/m³) H_2S for 20, 40, or 60 minutes. Heart rate, blood pressure, and electrocardiogram were monitored during exposure and for one hour postexposure. At 20, 40, and 60 minutes, heart rates decreased by 10, 17, and 27%, respectively, and continued to decrease by 29, 35, and 38%, respectively, during the postexposure monitoring period. Electrocardiograms revealed that the P-Q intervals were longer during the exposure and postexposure periods, which suggests some disorder in stimulus transmission. No changes were noted in blood pressure either during or after exposure.

4.4.4. Ocular Toxicity

Hydrogen sulfide has previously been reported to affect the eyes in a condition called "sore eye" or "gas eye". Lefebvre et al. (1991) examined this effect by exposing male Fischer 344 rats to 560 mg/m³ (403 ppm) H₂S for 4 hours or 2100 mg/m³ (1511 ppm) H₂S for 4 min. Ocular lavages were performed and cells were examined. There was a significant increase (approximately 2 times) in exfoliated ocular cells associated with both H₂S exposures. In all animals, the majority of recovered cells were corneal epithelial cells. However, a greater percentage of conjuctival epithelial cells was detected in exposed animals. The study demonstrates that H₂S is an ocular irritant at very high concentrations.

4.4.5. Genotoxicity

The mutagenic potential of H_2S was investigated using the Ames *Salmonella typhimurium* mutagenicity assay with and without Aroclor-induced hamster and rat liver S9 fractions. Hydrogen sulfide vapor (17-1750 µg/plate) was not mutagenic in *S. typhimurium* strains TA 97, TA 98, or TA 100 with and without metabolic activation (Hughes et al., 1984). Although H_2S is not mutagenic, Berglin and Carlsson (1986) reported that H_2S gas potentiated the mutagenicity of hydrogen peroxide in *S. typhimurium* strain TA 102. The investigators

believed that the increased mutagenicity was due to H_2S -mediated formation of iron sulfide, which converts hydrogen peroxide to hydroxyl radicals more efficiently than ferrous iron.

4.5. EVALUATION OF MODE OF ACTION FOR NONCANCER EFFECTS

4.5.1. Mode of Action of Noncancer Effects

At low levels of exposure H_2S is an irritant such that effects on tissues-of-contact would be anticipated, either in laboratory animals or humans (Shusterman, 1992).

At high levels, neurotoxic effects have been reported. Neurotoxicity is thought to occur via the ventilatory centers of the brain that lead to secondary effects including respiratory paralysis, pulmonary edema and death. Some investigators interpret the potential mode(s) of action for H2S very differently, whether it be via a neurotoxic pathway or effects on the respiratory system.

The potential mode-of-action of the neurotoxicity of H_2S has been investigated. Studies included neurochemical, neuroelectrophysiological, and behavioral approaches to address this issue. Baldelli et al. (1993) concluded from studies using H_2S with or without simultaneous ventilation that the neuronal necrosis seen following H_2S exposure is not a direct result of H_2S , but an effect of hypotension and ischemia. Warenycia et al. (1989a) demonstrated a selective uptake of sulfide by the brainstem of rats that, in part, may explain the lethal effect of sulfide on respiratory function as a consequence of an interaction with the brainstem respiratory center. Most of the catecholaminergic innervation of the brain originates from within the brainstem, while catecholamines and serotonin affect respiratory rhythm (Warenycia et al., 1989b). To explore this possibility further, experiments on monoamine oxidase (MAO) enzyme activity and neurotransmitter levels in various regions of the brain have been performed in rats (Warenycia et al., 1989b). It was concluded that the inhibition of MAO activities and increased neurotransmitter levels in the brainstem may be important in the loss of central respiratory drive after H_2S exposure.

On the other hand, some investigators believe that pulmonary edema might be a primary cause of death following exposure to high concentrations of H_2S (Lopez et al., 1989), possibly by alterations in pulmonary surfactants (Green et al., 1991). Other investigators, using dogs, have suggested that sulfide was an irritant to the afferent endings of the pulmonary vagi, and that high amounts of sulfide tend to paralyze the ventilatory center (Haggard and Henderson, 1922).

Almeida and Guidotti (1999), on the basis of rat studies, suggest that the lung may be the primary locus of HS ion action that can lead to vagal paralysis. Khan et al. (1990) investigated the action of H_2S on a molecular level in the lung, in which lung mitochondria incubated with H_2S exhibited a dose-dependent inhibition of cytochrome oxidase (CytOx) activity. Exposure to

50 ppm (70 mg/m³) H_2S and higher significantly reduced CytOx activity in lung mitochondria in a dose-dependent manner. Since CytOx is the terminal step in aerobic oxidative metabolism, inhibition by H_2S would stop tissue respiration and lead to the development of hypoxia.

Case reports involving individuals in accidental overexposure situations indicate that H₂S can cause direct neurotoxicity leading to respiratory distress and a variety of secondary adverse effects such as pulmonary edema, ataxia, motor weakness, headache, nausea, dyspnea, chest pain and dizziness. In some instances, death resulted from hypoxia. The concentrations causing these effects are largely unknown. Concentrations estimated to cause loss of smell are in the range of 100-200 ppm (140-280 mg/m³). Those associated with immediate physical collapse or "knockdown"have been estimated to be in the range of 500-1,000 ppm (695-1,390 mg/m³). Levels associated with secondary pulmonary edema have been estimated to be in the range of 250-500 ppm (348-695 mg/m³) (Milby and Baselt, 1999a,b; Reiffenstein et al., 1992).

Limited epidemiological studies (Richardson, 1995; Hessel et al., 1997; Vanhoorne et al., 1995), supported by findings in laboratory animals, have reported cardiovascular, pulmonary, and ocular effects attributable to H_2S . High concentrations of H_2S , such as in the Breysse (1961) report "as high as 4,000 ppm (5,560 mg/m³) H_2S ," can result in pulmonary edema and death. Chronic exposure to low concentrations on the order of 0.3 to 7.8 ppm (0.4 to 11 mg/m³) (Higashi et al., 1983) appears to result in no change in pulmonary function in humans.

Inhalation studies in laboratory animals (rat and mouse) have been subchronic in nature and have identified the olfactory nasal mucosa as the principal target site affected (Dorman et al., 2000; Brenneman et al., 2000), most likely as a consequence of the irritant properties of H₂S. In these two subchronic studies, nasal lesions of the olfactory mucosa were observed in animals exposed to 30 or 80 ppm (42 or 111 mg/m³) and consisted of olfactory neuron loss and basal cell hyperplasia. No effects were observed at 10 ppm (14 mg/m³). Male and female B6C3F1/CrlBr mice exhibited minimal to mild inflammation of the anterior portion of the nasal mucosa when exposed to 80 ppm (111 mg/m³) H₂S for 6 hr/day, 5 days/week for 90 days (CIIT, 1983a). Only minimal lesions were found in Sprague-Dawley or Fischer 344 rats (CIIT 1983 b,c). While strain and species differences may explain the discrepancy in response between the Sprague-Dawley rats and the Fischer-344 rats and B6C3F1 mice, it is not known why contrasting effects of H₂S on nasal lesions were seen by Dorman et al. (2000) vs. CIIT (1983c) since Sprague-Dawley rats were used in both cases. This discrepancy may relate to the use of computational fluid dynamics models (e.g., Kimbell and Subramaniam, 2001) to determine specifically where to look for lesions. This methodology was not available at the time of the earlier CIIT (1983a,b,c) studies.

It has been made clear that exposure to H_2S can elicit neurologic alterations in animals that are to be regarded as possible indicators of a neurotoxic effect. These alterations are elicited by exposure to H_2S near to those at which nasal lesions of Brenneman et al. (2000) are observed. Neurotransmitter levels were significantly altered in the brains of postpartum rats pups exposed *in utero* and postpartum to 28 mg/m³ (20 ppm) H₂S (Skrajny et al., 1992). Altered morphology of the cerebellar Purkinje cells was noted by Hannah and Roth (1991), also at 28 mg/m³ H₂S. Evidence for mechanisms by which neurotoxic effects may occur implicates ischemia mediated by H₂S inhibition of respiratory chain enzyme activity, in particular cytochrome oxidase. Khan et al. (1990) reported significant decreases in several respiratory chain enzymes during exposures at H₂S above but not at 10 ppm (14 mg/m³). This closeness in exposures eliciting these various effects indicates that this exposure range is in the critical transitional areas in the exposure-dose-severity continuum for a variety of endpoints, with data available in some instances to discriminate the mechanistic transition from being a nonadverse- to an adverse-response as are the olfactory lesions noted by Brenneman et al. (2000) at a NOAEL of 10 ppm (14 mg/m³).

4.5.2. Oral Exposure

In a subchronic oral study in rats, the only one that was found, mortality and neuromuscular and behavioral signs have been reported in animals gavaged with 7 mg/kg-day H_2S in deionized/deoxygenated water for 89 days (Anderson, 1987). Effects included labored breathing, convulsions, and sedation. Mortality was also noted in some males and females exposed to 3.5 mg/kg-day H_2S for 89 days. Although restlessness was observed in all treated males and increased salivation in all treated females, these clinical signs generally occurred sporadically, and the incidence of these findings did not show a clear dose-response. The toxicological significance of restlessness in treated males, and salivation in treated females, is not clear. No histopathological lesions attributable to H_2S were found. However, this study has serious problems with execution that prevent it from being used for derivation of an RfD. These problems include a prevalence of pneumonia and ulcerative dermatitis in vehicle control and treated groups, that fact that the study is not peer-reviewed, and that although 50% of the high dose males died, the cause of death could not be determined.

4.6. WEIGHT-OF-EVIDENCE EVALUATION AND CANCER CHARACTERIZATION—SYNTHESIS OF HUMAN, ANIMAL, AND OTHER SUPPORTING EVIDENCE, CONCLUSIONS ABOUT HUMAN CARCINOGENICITY, AND LIKELY MODE OF ACTION

Under the Draft Revised Guidelines for Carcinogen Risk Assessment (U.S. EPA, 1999), data are *inadequate* for an assessment of the carcinogenic potential of hydrogen sulfide. No human or animal studies assessing the potential carcinogenicity of H_2S were located. Genotoxicity studies also indicate that H_2S was not mutagenic in *S. typhimurium* strains TA97, TA98 or TA100 with or without metabolic activation (Hughes et al., 1984). Although H_2S does not appear to be mutagenic, H_2S gas potentiated the mutagenicity of hydrogen peroxide in *S. typhimurium* strain TA102 (Berglin and Carlsson, 1986). The investigators reported that the increased mutagenicity was due to H_2S -mediated formation of iron sulfide, which converts hydrogen peroxide to hydroxyl radicals more efficiently than ferrous iron alone.

4.7. SUSCEPTIBLE POPULATIONS

Data reported by Hannah and Roth (1991) showed that *in utero* and perinatal exposure of Sprague-Dawley rats to 20 or 50 ppm (28 or 70 mg/m³) H_2S resulted in abnormal growth and morphology of developing cerebellar Purkinje cells and changes in neurotransmitter levels (Skrajny et al., 1992; Hannah et al., 1989). Therefore pregnant women might be considered as a susceptible subpopulation with the hazard, if any, to the offspring of these women.

4.7.1. Possible Childhood Susceptibility

Conclusions about childhood susceptibility are extremely limited because of poor exposure data reported in existing case reports and lack of supporting data in the literature. There is some anecdotal evidence (following) that children may be more susceptible than adults to the adverse effects H_2S exposure.

A case study reported that a 20-month-old child developed intermittent paroxysmal tonic deviation of the eyes, choreoathetosis and dystonia with progressive involuntary movements of the entire body following exposure to at least 0.6 ppm (0.8 mg/m^3) H₂S over a period of up to "nearly a year" (Gaitonde et al., 1987). From the text within the report, coexposure to the same levels of H₂S can be inferred for the parents and at least one other child; this is evidence, although anecdotal, that children and possibly children of a certain age may be selectively susceptible to the neurological effects of H₂S at least at relatively high levels, > 600 ppb. Ascribing relevance of this apparent susceptibility to environmental levels of H₂S would be, however, conjectural.

Based on the findings of the effects of H_2S on developing brain cells in rats treated both pre- and postnatally (Hannah and Roth, 1991; see Section 4.3) and the fact that in humans the period of synaptogenesis extends from the sixth month of gestation to several years after birth (Ikonomidou et al., 2000), a possibility of children's vulnerability to effects from high levels of exposure to H_2S may exist.

4.7.2. Possible Gender Differences

Dorman et al. (2000) examined effects on reproductive performance from inhalation exposure to H_2S (see Section 4.3). There were no statistically significant effects on reproductive performance (mating index, fertility index, postimplantation loss per litter, and number of late

resorptions or stillbirths) in F_0 female animals. Also, the number of live pups, litter size, average length of gestation, and average number of implants per female were not affected. In comparing high-exposure and control male animals, a few histologic diagnoses related to testicular tubular degeneration and changes in the epididymis, with a higher incidence noted in the high-exposure group. In none of these cases, however, was statistical significance achieved.

5. DOSE-RESPONSE ASSESSMENTS

5.1. ORAL REFERENCE DOSE

Relevant quantitative human subchronic and chronic oral toxicity data are not available. Only one subchronic oral animal study was found (Anderson, 1987). A number of problems with that study precluded the derivation of an oral RfD. Among the problems with that study are the following: (1) it is unpublished and therefore not peer-reviewed, (2) a 75 to 100% incidence of pneumonia was reported in vehicle controls and all treatment groups. However, a 0% incidence of pneumonia was reported for the rack control group (which was not gavaged). Major problems were evident in the execution of this study. The prevalence of pneumonia across all gavaged groups indicates a prevalence of disease that precludes drawing conclusions from this study regarding the effect of H_2S , (3) there was a 50% death rate in the high dose male group, the only group with significant mortality, but the cause of death "could not be determined." This also inspired a lack of confidence in the data, (4) alopecia was common in all groups, and the problem was diagnosed by a veterinarian as ulcerative dermatitis. This disease plus the pneumonia prevents one from calling this group "normal" and thus appropriate for the production of reliable data.

5.2. INHALATION REFERENCE DOSE

5.2.1. Choice of Principal Study and Critical Effect with Rationale and Justification

It is clear from the clinical findings of numerous case reports of overexposure situations and occupational studies that neurotoxicity is the principal concern at high exposure levels of H_2S . Several groups of investigators (Tvedt et al., 1991a,b; Kilburn, 1993, 1997; Wasch et al., 1989) have reported long-term adverse neurological sequelae of H_2S -induced unconsciousness in humans during occupational, accidental, and chronic exposures including neuropsychological and neurobehavorial decrements, brain damage, and, in one instance, dementia.

As reported and analyzed in prior Sections (see 4.3.3, 4.4.1, and 4.5.2), both nasal tract lesions and neurologic effects occur in the same concentration range at lower levels of H_2S

exposure, and both are clearly of relevance to humans. Based on the relative clarity of the mechanistic basis, the dose-response information at lower levels of H_2S exposure, and adverse nature of the endpoint for nasal tract lesions as compared to the neurologic endpoints, the former is chosen as the critical effect on which to base the noncancer dose-response assessments. That the levels of H_2S associated with these effects appear to be similar for either endpoint (e.g., the possible indicator of neurotoxic effects reported by Hannah and Roth (1991) to occur at 28 mg H_2S/m^3 and the no-effect level of nasal tract lesions reported by Brenneman et al. (2000) at 14 mg H_2S/m^3) is some indication that consideration for the one will also address the other. On the other hand, the same judgments made here concerning clarity of mechanism or dose-response or adversity could serve as the basis for a reevaluation of the critical endpoint if these matters were further resolved for neurologic effects.

The Purkinje-cell alterations noted by Hannah and Roth (1991) are a chemically induced change in the structure of the nervous system. While such alterations may constitute a possible indicator of a neurotoxic effect defined in the U.S. EPA neurotoxicity risk assessment guidelines as "neuroanatomical changes occurring at any level of nervous system organization" (U.S. EPA, 1998), their biological significance is not sufficiently understood to clearly ascertain the nature of the adverse outcome. The effects of certain developmental insults on cerebellar structure and function have been well characterized (Ferguson 1996) and provide a basis with which to gauge the significance of the Hannah and Roth (1991) findings. Focal x-irradiation of cerebella of rats during the first postnatal week (before PND8 which is inclusive of the exposure period used by Hannah and Roth, 1991) reduces the number of granule, basket and Purkinje cells, and surviving Purkinje cells are misaligned and exhibit aberrant dendritic form. These animals exhibit tremor, ataxia, altered spontaneous activity and their performance on rotorod and cognitive tests (swim maze) is impaired. Similar neuroanatomical and functional alterations are observed after early life treatment of rats (PND 4 onwards) with antimitotic agents such as methylazoxymethanol. Dorman et al., (2002) did not find evidence of the types of neuroanatomical alterations or behavioral impairments associated with these models in their dose-response assessment of developmental exposure to hydrogen sulfide. While it is noteworthy that Dorman et al. (2002) did not did not assess Purkinje cell dendritic morphology or employ a cognitive test sensitive to cerebellar insult, such as conditioned eye-blink, their observations indicate that developmental exposure to hydrogen sulfide does not produce the substantial neuroanatomical or behavioral alterations associated with well-characterized models of developmental cerebellar insult. While a definitive study with hydrogen sulfide in which neural structure and function are intensively examined is not available, the currently available data do not indicate that developmental hydrogen sulfide exposure is associated with overt pathology or functional deficits. Furthermore, the currently available evidence does not provide a basis to clearly ascertain adverse consequences associated with the subtle alterations in Purkinje cell dendrites reported by Hannah and Roth (1991). Nonetheless, it should be noted that the exposure levels used by Hannah and

Roth (1991) are nearly the same as those used in eliciting the nasal tract effects discussed below, indicating that the uncertainty factors applied to the nasal effects would also provide a similar degree of conservatism for both of these endpoints.

The few available occupational studies in humans are judged not to be suitable for derivation of an inhalation RfC because of limitations including lack of adequate exposure data, limited number of subjects examined, and co-exposure to other chemicals and lack of information on the critical effect (nasal toxicity) designated above.

Three subchronic animal studies were considered for derivation of an inhalation RfC: Dorman et al. (2000), CIIT (1983a), and Brenneman et al. (2000). In comparing high-exposure and control male animals exposed subchronically to 80 ppm H₂S Dorman et al. (2000) reported a few histologic testicular diagnoses in the high-exposure group (80 ppm); in none of these cases, however, was statistical significance achieved when compared to the control group. CIIT (1983a) reported that male and female mice exposed to 80 ppm (111 mg/mg³) H₂S, 6 hr/day, 5 days/week for 90 days exhibited minimal to mild inflammation of the anterior portion of the nasal mucosa. Only minimal lesions were observed in two species of rats (Sprague-Dawley and Fischer-344) examined by the same authors (CIIT, 1983b,c) and in mice exposed to lower concentrations of H₂S. However, in male Sprague- Dawley rats exposed to 0, 10, 30, and 80 ppm (0, 14, 42, and 111 mg/m³) H_2S for 6 hr/day, 7 days/week, for 10 weeks, Brenneman et al. (2000) reported nasal lesions of the olfactory mucosa in the 30 and 80 ppm (42 and 111 mg/m³) exposure groups. No effects were observed in the control or 10 ppm (14 mg/m^3) exposure animals that were considered treatment-related. Nasal lesions consisted of multifocal, bilaterally symmetrical olfactory neuron loss and basal cell hyperplasia affecting the lining of the dorsal medial meatus and dorsal and medial region of the ethmoid recess. The difference in results between Brenneman et al. (2000) and CIIT (1983a) may indicate a variability in animal response.

The study by Brenneman et al. (2000) was considered to be the most appropriate for the derivation of an inhalation RfC for several reasons. First, the critical effect (nasal lesions of the olfactory mucosa; summarized in Table 2 in Section 4.2.2) has been reported by other investigators (Dorman et al., 2000; CIIT, 1983a; Lopez et al., 1988b); second, the effect is consistent with the irritant properties of this gas; third, along with the neurological system, the respiratory system has been reported be a target organ of H_2S toxicity by numerous researchers; fourth, the LOAEL (42 mg/m³) and NOAEL (14 mg/m³) are at lower concentrations than those in the other subchronic studies.

5.2.2. Methods of Analysis—including Models (PBPK, BMD, etc.)

The results of Brenneman et al. (2000) (Table 2, Section 4.2.2) yielded a NOAEL of 10 ppm and a LOAEL of 30 ppm. The nature of the data (total incidence preceded by zero

incidence; see Table 2) was not amenable to BMD modeling, thus NOAEL and LOAELs were designated. Converting ppm to mg/m³ resulted in a NOAEL of 14 mg/m³ and a LOAEL of 42 mg/m³ (1 ppm = 1.39 mg/m³). The LOAEL and NOAEL were then adjusted to a weekly continuous exposure such that (N)LOAEL_{ADJ} = (N)LOAEL (mg/m³) x hrs exp/24 hours x day exp/week :

NOAEL_{ADJ} = 14 mg/m³ x 6/24 hours x 7/7 days/week = 3.5 mg/m^3 LOAEL_{ADJ} = 42 mg/m^3 x 6/24 hours x 7/7 days/week = 10.5 mg/m^3

Thus, a $LOAEL_{ADJ}$ of 10.5 mg/m³ and a $NOAEL_{ADJ}$ of 3.5 mg/m³ were obtained.

Conversion of these (adjusted) effect levels to human equivalent concentrations, i.e. HECs, is typically accomplished through application of procedures specified in the inhalation RfC methodology (U.S. EPA 1994b). Species specific parameters of anatomy, physiology and location of the lesion are utilized to generate a dosimetric adjustment factor (DAF) that is then applied to the effect level of interest, a NOAEL in this instance, to generate a NOAEL_{HEC} For H₂S, this DAF is termed as the "regional gas dose ratio" (RGDR) and is calculated as the ratio of rat to human ventilation rate over the surface area of the ET region (see U.S. EPA, Chapter 4). The ventilation rate (V_E) was calculated for male Sprague-Dawley rats using the default body weight of 0.267 kilograms shown in Table 4-5 of U.S. EPA (1994b). The ventilation rate for male rats was calculated as follows:

$$\ln V_{\rm E} = b_0 + b_1 \ln(0.267 \text{kg})$$

The parameters, b_0 and b_1 , are species specific parameters listed in Table 4-6 of U.S. EPA (1994). This gave a V_E of 0.19 liters/minute for male rats. The default ventilation rate for humans is 13.8 liters/minutes given on page 4-27 of U.S. EPA (1994b). The default ET surface areas for the rat and human are 15 and 200 cm², respectively. These values are given on page 4-26 of U.S. EPA (1994b). The RGDR was calculated as follows:

$$RGDR_{ET} = (V_E/SA_{ET})_{animal}/(V_E/SA_{ET})_{human} = (0.19/15)/(13.8/200) = 0.184$$

The animal (N)LOAEL_{ADJ} is then multiplied by the $RGDR_{ET}$ to yield the (N)LOAEL_{HEC}.

$$NOAEL_{HEC} = NOAEL_{ADJ} \times RGDR_{ET} = 3.5 \text{ mg/m}^3 \times 0.184 = 0.64 \text{ mg/m}^3$$

 $LOAEL_{HEC} = LOAEL_{ADJ} \times RGDR_{ET} = 10.5 \text{ mg/m}^3 \times 0.184 = 1.9 \text{ mg/m}^3$

The computational fluid dynamic (CFD) modeling information provided by Moulin et al. (2002) on the nasal pathology of H_2S provide useful insight into the results from the current default RfC procedure used in constructing the HEC for H_2S . From the work of Brenneman et al. (2000) it is clear that the lesions observed in rats from H_2S exposure are restricted to the olfactory epithelium and do not impact the respiratory epithelium. Moulin et al. (2002) further refined these findings by showing that CFD simulated levels of H_2S flux (and therefore H_2S dose) correlated with the incidence of lesions within the olfactory epithelium, suggesting that the appropriate dose response metric for H_2S would be highly specific, i.e., flux to the olfactory epithelium. The current RfC default procedures (used above) do not employ any kind of CFD-type air flow adjustment, nor do they use a surface area measure that discriminates between olfactory and respiratory epithelium.

Thus, although the modeling information provided by Moulin et al. (2002) is not used in this assessment to provide quantitative assessment of species comparison of H_2S exposure (i.e, an HEC) the results may be used to provide valuable insight into the limitations of the current procedures. Future work in this area may provide information to perform a more quantitative evaluation of the HEC for H_2S .

5.2.3. RfC Derivation—including Application of Uncertainty Factors (UFs) and Modifying Factors (MFs)

The RfC for H_2S is derived by dividing the NOAEL_{HEC} for nasal effects by the UF of 300. Uncertainty factors were applied for interspecies extrapolation from animals to humans (3) and intraspecies variability (10). The interspecies uncertainty factor of 3 (10^{1/2}) was used rather than 10 because of a dosimetric adjustment from rat to human was used. The use of an additional uncertainty factor of 10 was employed for the use of a subchronic, rather than chronic, study.

Although a single generation reproductive study noted testicular alterations that were observed only in the high dose group, the alterations were not significantly different from controls and had no apparent effects on reproductive performance. No other indicators of reproductive toxicity were observed in this study. No significant histopathology of reproductive organs was noted in a longer duration (subchronic) study. These results can be considered to lessen the concern for lack of a multi-generational reproductive study.

As noted above, the significance and relevance of the Purkinje dendritic alterations noted in developing animals are indicated in the neurotoxicity risk assessment guidelines to be a possible indicator of a neurotoxic effect, not as a clearly adverse effect. The RfC is based on an effect that could be considered a neurological effect (olfactory neuronal loss) that is clearly adverse occurring in adult animals at a lower dose than did the altered dendritic arborization in developing animals. This dose-response relationship between adult and neonatal effects tends to ameliorate concerns that children are a susceptible population at least in the situation of longterm relatively low level exposures. Some circumstantial data seem to indicate that, under high level exposures, children may be more susceptible than adults to the acute effects of hydrogen sulfide.

In consideration of the points above and the totality of information available on hydrogen sulfide, no data base uncertainty factor is evoked.

Therefore, a total uncertainty factor of 300 (10 for subchronic to chronic, 10 for intraspecies variability, and 3 for interspecies extrapolation) is applied to the NOAEL_{HEC} of 0.64 mg/m³, yielding an inhalation RfC of 0.002 mg/m³.

 $RfC = NOAEL_{HEC} / Uncertainty Factors$ $RfC = 0.64 \text{ mg/m}^3 / (10 \text{ x } 10 \text{ x } 3) = 0.002 \text{ mg/m}^3 = 2E-3 \text{ mg/m}^3$

5.3. CANCER ASSESSMENT

No data pertaining to the potential carcinogenicity of H₂S were identified.

6. MAJOR CONCLUSIONS IN THE CHARACTERIZATION OF HAZARD AND DOSE RESPONSE

6.1. HAZARD IDENTIFICATION

Hydrogen sulfide is a colorless gas and has a strong odor of rotten eggs. Its primary uses include the production of elemental sulfur and sulfuric acid, the manufacture of heavy water and other chemicals, in metallurgy, and as an analytical reagent. Although quantitative data are lacking, toxicity studies suggest that H_2S gas is absorbed rapidly through the lungs. Oral exposure is not likely to occur. In animals and humans, it distributes to the blood, brain, lung, heart, liver, spleen, and kidney. Oxidation is the primary metabolic pathway for H_2S , with thiosulfate and sulfate as metabolites. Metabolism in laboratory animals and in humans appears to be similar. Hydrogen sulfide is excreted in the urine.

Human data pertaining to inhalation exposure (the expected route of ambient exposure) consist of a plethora of case reports and a variety of occupational epidemiological studies. Although these studies have limitations that preclude their use for quantitative risk assessment, they indicate that exposure to H_2S (at high concentrations) has profound effects on the respiratory system leading to unconsciousness with attendant neurologic sequelae and, sometimes, death. An increase in cardiovascular-related deaths due, in part, to H_2S exposure was reported in one occupational study.

Inhalation studies in adult rodents demonstrate sensitivity of nasal olfactory epithelium to low concentrations of H_2S . The RfC is based on these lesions. Limited evidence suggests that exposure of humans to low concentrations may also cause neurologic symptoms although quantitative exposure-response data is lacking. Because of similar access of inhaled H_2S to the olfactory tissues of humans, these lesions are likely of relevance to humans and a reasonable choice as a critical effect. Relevance to olfactory lesions seen in rodents to humans is also suggested by Hirsch and Zavala (1999) who reported decreased persistent olfactory function in workers exposed to hydrogen sulfide chronically. Whereas adverse nasal effects are of relevance and concern in adult human exposure scenarios, inhalation studies of perinatal or neonatal exposure in rats demonstrates abnormal cellular development in the brain as well as significant alterations in neurotransmitter levels; the toxicological significance of these findings is uncertain.

Relevant quantitative human oral toxicity data are not available and ingestion is not a likely route of exposure. An RfD based on GI disturbances in pigs consuming feed containing hydrogen sulfide was derived in the previous IRIS entry. A review of the RfD (see Section 4.2.1) indicates that the effects on which that value were based are not reproducible and probably not related to H_2S . Therefore the previous RfD will be withdrawn and a new RfD will not be derived based on data base deficiencies.

The indicators of a possible effect noted in the developing brain cells of newborn rats indicate the possibility that the developing human fetus could also be at risk. The exposure levels producing these effects, however, are in the same range or somewhat higher than those producing the critical sentinel clearly adverse effect (nasal tract lesions) in adult animals, thereby ameliorating the concern that young animals (and possibly children) may be especially susceptible to the effects from relatively low-level chronic exposures to hydrogen sulfide. Other observations in the data base do indicate a possible concern regarding the susceptibility of children exposed to high levels of H_2S , i.e., > 600 ppb. However, the relevance of this apparent susceptibility at environmental levels of H_2S where toxicity is not likely to occur, such as the RfC value derived herein, is not at all clear.

There is no evidence indicating that H_2S exposure is associated with carcinogenesis. Under the Draft Revised Guidelines for Carcinogen Risk Assessment (U.S. EPA, 1999), data are *inadequate* for an assessment of the carcinogenic potential of hydrogen sulfide.

6.2. DOSE RESPONSE

The effects of H_2S in humans can be acute and/or chronic. The weight-of-evidence from the human and animal studies suggest that the effects of H_2S are dose-dependent. The exposure-response relationship for acute effects, particularly CNS and respiratory, can be very steep. Levels in the range of 500 to 1,000 ppm (695 to 1,390 mg/m³) are life-threatening and can cause

immediate unconsciousness followed by serious and debilitating neurologic and respiratory sequelae. Lower levels have been associated with lung function deficit and eye, nose, and throat irritation. Precise determination of acute effect levels is precluded by the lack of accurate monitoring data since nearly all cases were those involving acute accidental overexposure. Levels dangerous to human health may not be detected by odor since high levels of H₂S can paralyze the olfactory nerves making detection impossible.

Little is known of the concentration-response relationship at low levels of chronic exposure. Occupational studies have been confounded by exposure to other substances and inadequate monitoring to establish cause-effect relationships.

Controlled exposure studies in laboratory animals indicate that the portal-of-entry (the nasal tract) and neurologic tissues are both targets for H_2S toxicity. Based on the clear nature of their adversity (demonstrable histopathology), information available on exposure-dose-response and the commonality of the underlying mechanism (irritation) between animals and humans, and indications that such effects also occur in humans chronically exposed to hydrogen sulfide (Hirsch and Zavala, 1999), nasal tract effects were chosen as the critical effect and used to develop an inhalation reference concentration (RfC). Available information indicates that the dose-response character of indicators of neurotoxicity (such as the alterations to cerebellar Purkinje cells reported by Hannah and Roth, 1991) and nasal tract lesions (Brenneman et al., 2000) may be similar to one another such that consideration of one may be inclusive of the other. However, more extensive and definitive information on the neurologic endpoints could reveal that these should be the most relevant endpoint, more so than nasal tract lesions. Such information may provide sufficient reason to reassess H_2S .

Some data exist inferring that children/neonates may be more susceptible than adults to the neurological effects of H_2S . In the case-study of Gaitonde et al. (1987) neurological effects observed in a 20-month-old that were associated with an extended exposure to high levels of H_2S (>600 ppb) were not reported in the parents or other children in the family. It is clear that the findings of Hannah and Roth (1991) on developing brain cells in rats exposed both pre- and postnatally to high levels of H_2S (20 and 50 ppm) are possible indicators of a neurotoxic effect. However, the toxicologic significance of the observed alterations are not yet understood although the effects noted in both of these reports possibly having the same underlying mechanism, inhibition of mitochondrial cytochrome oxidase activity. Overall, these observations indicate a possible concern regarding the susceptibility of children exposed to high levels of H_2S , i.e., > 600 ppb. However, the relevance of this apparent susceptibility at environmental levels of H_2S where toxicity is not likely to occur, such as the RfC value derived herein, is not at all clear.

The daily inhalation exposure to the human population that is likely to be without an appreciable risk of deleterious effects during a lifetime, the RfC, has been determined to be

0.002 mg/m³ or 1.4 ppb. This value was derived from a subchronic inhalation study in laboratory rats by Brenneman et al. (2000). This study reported critical effects, nasal tract lesions in olfactory epithelium, in male rats exposed to 30 or 80 ppm (42 or 111 mg/m³) but not at 10 ppm (14 mg/m³), the NOAEL. An overall uncertainty factor of 300 was applied to account for interspecies extrapolation from animal to human, intrapecies variability, and subchronic exposures. The use of nasal lesions permits the derivation of an RfC that is considered protective against any long-term CNS and respiratory effects in ambient exposure environments.

The overall confidence in the RfC assessment is medium-high; confidence in the principal study is medium because it was well-designed and conducted, and examined sensitive endpoints. However, it was subchronic in duration, and examined only male rats. Overall confidence in the data base is medium-high because the endpoints are supported by other subchronic studies. A chronic inhalation study in rats and mice, with additional dose groups and a more extensive comparative examination of neurologic and nasal tract endpoints from the same exposures, would increase the confidence in the data base. Ongoing scientific efforts to provide further refinements to the assessment of H_2S , particularly in the area of regional dosimetry of the nasal tract, are acknowledged.

7. REFERENCES

Allyn, LB. (1931) Notes on hydrogen sulfide poisoning. Ind. and Eng. Chem. 23:234.

Almeida, AF; Guidotti, TL. (1999) Differential sensitivity of lung and brain to sulfide exposure: a peripheral mechanism for apnea. Toxicol. Sci. 50:287-293.

Anderson, RC. (1987) 90-Day subchronic oral toxicity in rats. Test material: hydrogen sulfide (H_2S). Arthur D. Little, Inc. Report to Dynamac Corporation. Work Assignment Nol 11148-008. Reference 55461.

Andrew, FD, Renne, RA; Cannon, WC. (1980) Reproductive toxicity testing for effects of H_2S in rats. Pacific northwest laboratory annual report for 1979 to the DOE assistant secretary for environment: part 1 biomedical sciences. Contract Number EY-76-C-06-1830

Arnold, IMF; Dufresne, RM; Alleyne, BC; et al. (1985) Health implications of occupational exposures to hydrogen sulfide. J. Occup. Med. 27:373-376.

Baldelli, RJ; Green, FH; Auer, RN. (1993) Sulfide toxicity: mechanical ventilation and hypotension determine survival rate and brain necrosis. J. Appl. Physiol. 75(3):1348-1353.

Bartholomew, TC; Powell, GM; Dodgson, KS; et al. (1980) Oxidation of sodium sulphide by rat liver, lungs and kidney. Biochem. Pharmacol. 29:2431-2437.

Beauchamp, RO Jr; Bus, JS; Popp, JA; et al. (1984) A critical review of the literature on hydrogen sulfide toxicity. Crit. Rev. Toxicol. 13:25-97.

Berglin, EH; Carlsson, J. (1986) Effect of hydrogen sulfide on the mutagenicity of hydrogen peroxide in *Salmonella typhimurium* strain TA102. Mutat. Res. 175(1):5-9.

Berry, M; Flinn, R. (1984) Vertex analysis of Purkinje cell dendritic trees in the cerebellum of the rat. Proc. R. Soc. Lond. B 221 (321-348).

Borda, RP. (1997) Cognitive impairment associated with chronic low-level exposure to hydrogen sulfide. Presented at the American Public Health Association 125th annual meeting, November.

Brenneman, KA; James, RA; Gross, EA; Dorman, DC. (2000) Olfactory neuron loss in adult male CD rats following subchronic inhalation exposure to hydrogen sulfide. Toxicol. Pathol. 28(2):326-333.

Brenneman KA; Meleason DF; Sar M; Marshall MW; James RA; Gross EA; Martin JT; Dorman DC.(2002). Olfactory mucosal necrosis in male CD rats following acute inhalation exposure to hydrogen sulfide: reversibility and the possible role of regional metabolism. Toxicol. Pathol. 30(2):200-208

Breysse, PA. (1961) Hydrogen sulfide fatality in a poultry feather fertilizer plant. Industrial Hygiene Journal 22:220-222.

Burnett, WW; King, EG; Grace, M.; Eng, P; Hall, WF. (1977) Hydrogen sulfide poisoning: review of 5 years' experience. Canadian Med. Assn. J. 117:1277-1280.

CIIT (Chemical Industry Institute of Toxicology). (1983a) 90-Day vapor inhalation toxicity study of hydrogen sulfide in B6C3F1 mice. EPA/OTS 0883-0255.

CIIT (1983b) 90-Day vapor inhalation toxicity study of hydrogen sulfide in Fischer-344 rats. EPA/OTS 0883-0255.

CIIT (1983c) 90-Day vapor inhalation toxicity study of hydrogen sulfide in Sprague-Dawley rats. EPA/OTS 0883-0255.

Dorman, DC; Brenneman, KA; Struve, MF; et al. (2000) Fertility and developmental neurotoxicity effects of inhaled hydrogen sulfide in Sprague-Dawley rats. Neurotoxicol Teratol. 22:71-84.

Dorman, DC; Moulin, FJM; McManus, BE; et al. (2002) Cytochrome oxidase inhibition induced by acute hydrogen sulfide inhalation: correlation with tissue sulfide concentrations in the rat brain, liver, lung, and nasal epithelium. Toxicol. Sci. 65:18-25.

Elovaara, E; Tossavinen, A; Savolainen, H. (1978) Effects of subclinical hydrogen sulfide intoxication on mouse brain protein metabolism. Exp. Neurol. 62(1):93-98.

Ferguson, SA. (1996). Neuroanatomical and functional alterations resulting from early postnatal cerebellar insults in rodents. Pharmacol. Biochem. Behavior 55(4): 663-671.

Gaitonde, UB; Sellar, RJ; O'Hare, AE. (1987) Long term exposure to hydrogen sulfide producing subacute encephalopathy in a child. Br. Med. J. 294(6572):614.

Glass, DC. (1990) A review of the health effects of hydrogen sulphide exposure. Ann. Occup. Hyg. 34: 323-327.

Green, FH; Schèurch, S; De Sanctis, GT; et al. (1991) Effects of hydrogen sulfide exposure on surface properties of lung surfactant. J. Appl. Physiol. 70(5):1943-1949.

Guidotti, TL. (1996) Hydrogen sulfide. Occup. Med. 46(5): 367-371.

Gunina, AI. (1959) Oxidation of hydrogen sulfide in the organism. U.S. Atomic Energy Commission, Report No. AEC tr 3661:534-541.

Haahtela, T; Marttila, O; Vilka, V; et al. (1992) The South Karelia air pollution study: acute health effects of malodorous sulfur pollutants released by a pulp mill. Am. J. Public Health 82(4): 603-605.

Haggard, HW. (1921) The fate of sulfides in the blood. J. Biol. Chem. 49:519-529.

Haggard, HW; Henderson, Y. (1922) The influence of hydrogen sulphide upon respiration. Am. J. Physiol. 41:289-297.

Hagley, SR; South, DL. (1983) Fatal inhalation of liquid manure gas. Med. J. Aust. 2:459-460.

Hall, AH; Rumack, BH. (1997) Hydrogen sulfide poisoning: an antidotal role for sodium nitrite? Vet. Human Toxicol. 39(3): 152-154.

Hannah, RS; Roth, SH. (1991) Chronic exposure to low concentrations of hydrogen sulfide produces abnormal growth in developing Purkinje cells. Neurosci. Lett. 122: 225-228.

Hannah, RS; Hayden, LJ; Roth, SH. (1989) Hydrogen sulfide exposure alters the amino acid content in developing rat CNS. Neurosci. Lett. 99: 323-327.

Haskell Laboratory. (1994) The effects of chronic exposure to sulfonated hydrogen; the effect of chronic exposure to hydrogen sulfide (and with carbon disulfide, with cover letter dated 05/11/94). EPA/OTS (1339). Doc. ID 86940000854.

Hayden, LJ; Goeden, H; Roth, SH. (1990a) Exposure to low levels of hydrogen sulfide elevates circulating glucose in maternal rats. J. Toxicol. Environ. Health 31(1):45-52.

Hayden, LJ; Goeden, H; Roth, SH. (1990b) Growth and development in the rat during subchronic exposure to low levels of hydrogen sulfide. Toxicol. Ind. Health 6(3):389-401.

Hessel, PA; Herbert, FA; Melenka, LS; et al. (1997) Lung health in relation to hydrogen sulfide exposure in oil and gas workers in Alberta, Canada. Am. J. Ind. Med. 31(5):554-557.

Higashi, T; Toyama, T; Sakurai, H; et al. (1983) Cross-sectional study of respiratory symptoms and pulmonary functions in rayon textile workers with special reference to H2S exposure. Ind. Health 21(4):281-292.

Higuchi, Y; Fukamachi, M. (1977) Behavioral studies on toxicity of hydrogen sulfide by means of conditioned avoidance responses in rats. Folia Pharmacol. Jpn. 73:307-319.

Hirsch, AR; Zavala, G. (1999) Long-term effect on the olfactory system of exposure to hydrogen sulphide. Occup. Environ. Med. 56: 284-287.

Hughes, TJ; Sparacino, C; Frazier, S. (1984) Validation of chemical and biological techniques for evaluation of vapors in ambient air testing of twelve (12) vapor-phase compounds (project summary). Research Triangle Park, NC: Environmental Protection Agency, Health Effects Research Laboratory 3:EPA-6001.

Ikonomidou, C; Bittigau, P; Ishimaru, MJ; et al. (2000) Ethanol-induced apoptotic neurodegeneration and fetal alcohol syndrome. Science 287(5455):1056-1060

Imamura, T; Kage, S; Kudo, K; et al. (1996) A case of drowning linked to ingested sulfides: a report with animal experiments. Int. J. Legal Med. 109(1):42-44.

Kage, S; Nagata, T; Kimura, K; et al. (1992) The usefulness of thiosulfate as an indicator of hydrogen sulfide poisoning in forensic toxicological examination: a study with animal experiments. Jpn. J. Forensic Toxicol. 10(3):223-227.

Kangas, J; Savolainen, H. (1987) Urinary thiosulphate as an indicator of exposure to hydrogen sulphide vapour. Clinica. Chimica. Acta. 164:7-10.

Khan, AA; Schuler, MM; Prior, MG; et al. (1990) Effects of hydrogen sulfide exposure on lung mitochondrial respiratory chain enzymes in rats. Toxicol. Appl. Pharmacol. 103(3):482-490.

Khan, AA; Yong, S; Prior, MG; et al. (1991) Cytotoxic effects of hydrogen sulfide on pulmonary alveolar macrophages in rats. J. Toxicol. Environ. Health 33(1):57-64.

Kilburn, KH. (1993) Case report: profound neurobehavioral deficits in an oil field worker overcome by hydrogen sulfide. Am. J. Med. Sci. 306(5):301-305.

Kilburn, KH; Warshaw, RH. (1995) Hydrogen sulfide and reduced-sulfur gases adversely affect neurophysiological functions. Toxicol. Environ. Health 11(2):185-197.

Kilburn, KH. (1997) Exposure to reduced sulfur gases impairs neurobehavioral function. Southern Med. J. 90(10): 997-1006.

Kilburn, KH. (1999) Evaluating health effects from exposure to hydrogen sulfide: central nervous system dysfunction. Environ. Epidemiol. Toxicol. 1: 207-216.

Kimbell, JS; Godo, MN; Gross, EA; Joyner, DR; Richardson, RB; Morgan, KT. (1997) Computer simulation of inspiratory airflow in all regions of the Fisher 344 rat nasal passages. Toxicol. Appl. Pharmacol. 145: 388-398.

Kimbell, JS; Subramaniam, RP. (2001) use of computational fluid dynamics models for the dosimetry of inhaled gases in the nasal passages. Inhal. Toxicol. 13 (5): 325-334.

Kimura, K; Hasegawa, M; Matsubara, K; et al. (1994) A fatal disaster case based on exposure to hydrogen sulfide - an estimation of the hydrogen sulfide concentration at the scene. Foren. Sci. Int. 66:111-116.

Kleinfeld, M; Giel, C; Rosso, A. (1964) Acute hydrogen sulfide intoxication: an unusual source of exposure. Ind. Med. Surg. 33:656-660.

Kohno, M; Tanaka, E; Nakamura, T; et al. (1991) Influence of the short-term inhalation of hydrogen sulfide in rats. Eisei. Kagaku 37(2):103-106.

Kombian, SB; Warenycia, MW; Mele, FG; et al. (1988) Effects of acute intoxication with hydrogen sulfide on central amino acid transmitter systems. Neurotoxicology 9(4):587-595.

Kosmider, S; Rogala, E; Pacholek, A. (1967) Electrocardiographic and histochemical studies of the heart muscle in acute experimental hydrogen sulfide poisoning. Arch. Immun. Ther. Exp. 15:731-740.

Laug, EP; Draize, JH. (1942) The percutaneous absorption of ammonium hydrogen sulfide and hydrogen sulfide. J Pharma col Exp Ther 76:179-188.

Lefebvre, M; Yee, D; Fritz, D; et al. (1991) Objective measures of ocular irritation as a consequence of hydrogen sulphide exposure. Vet. Hum. Toxicol. 33(6):564-566.

Legator, MS; Singleton, CR; Morris, DL; Philips DL. (2001) Health effects from chronic low-level exposure to hydrogen sulfide. Arch. Environ. Health. 56(2):123-31.

Levitt, MD; Furne, J; Springfield, J; et al. (1999) Detoxification of hydrogen sulfide and methanethiol in the cecal mucosa. J. Clin. Invest. 104(8):1107-1114.

Lopez, A; Prior, M ; Lillie, LE; et al. (1988a) Histologic and ultrastructural alterations in lungs of rats exposed to sub-lethal concentrations of hydrogen sulfide. Vet. Pathol. 25(5):376-384.

Lopez, A; Prior, M; Reiffenstein, RJ; et al. (1989) Peracute toxic effects of inhaled hydrogen sulfide and injected sodium hydrosulfide on the lungs of rats. Fundam. Appl. Toxicol. 12(2):367-373.

Lopez, A; Prior, M; Yong, S; et al. (1987) Biochemical and cytologic alterations in the respiratory tract of rats exposed for 4 hours to hydrogen sulfide. Fundam. Appl. Toxicol. 9(4):753-762.

Lopez, A; Prior, M; Yong, S; et al. (1988b) Nasal lesions in rats exposed to hydrogen sulfide for four hours. Am. J. Vet. Res. 49(7):1107-1111.

Mehlman, MA. (1994) Dangerous and cancer-causing properties of products and chemicals in the oil refining and petrochemical industry. Part VII: adverse health effects and toxic manifestations caused by exposure to hydrogen sulfide, a component of crude oil. In: Advances in Modern Environmental Toxicology, Vol XXIII. The identification and control of environmental and occupational diseases: hazards and risks of chemicals in the oil refining industry.

Milby, TH. (1962) Hydrogen sulfide intoxication; review of the literature and report of an unusual accident resulting in two cases of nonfatal poisoning. J. Occup. Med. 4: 431-437.

Milby, TH; Baselt, RC. (1999a) Health hazards of hydrogen sulfide: current status and future directions. Environ. Epidemiol. Toxicol. 1: 262-269.

Milby, TH; Baselt, RC. (1999b) Hydrogen sulfide poisoning: Clarification of some controversial issues. Am. J. Ind. Med. 35: 192-195.

Moulin, FJ; Brenneman, KA; Kimbell, JS; Dorman, DC. (2002) Predicted regional flux of hydrogen sulfide correlates with distribution of nasal olfactory lesions in rats. Toxicol. Sci. 66: 7-15.

National Research Council (NRC). (1983) Risk assessment in the federal government: managing the process. Committee on the Institutional means for Assessment of Risks to Public Health, Commission on Life Sciences. Washington, DC: National Academy Press.

Osbern, LN; Crapo, RO. (1981) Dung lung: a report of toxic exposure to liquid manure. Ann. Int. Med. 95:312-314.

Parra, O; Monso, E; Gallego, M; et al. (1991) Inhalation of hydrogen sulphide: a case of subacute manifestations and long term sequelae. Brit. J. Ind. Med. 48:286-287.

Partlo, LA; Sainsbury, RS; Roth, SH. (2001) Effects of repeated hydrogen sulphide (H_2S) exposure on learning and memory in the adult rat. Neurotox. 22:177-189.

Prior, MG; Sharma, AK; Yong, S; et al. (1988) Concentration-time interactions in hydrogen sulphide toxicity in rats. Can. J. Vet. Res. 52: 375-379.

Reiffenstein, RJ; Hulbert, WC; Roth, SH. (1992) Toxicology of hydrogen sulfide. Ann. Rev. Pharmacol. Toxicol. 109-134.

Richardson, DB. (1995) Respiratory effects of chronic hydrogen sulfide exposure. Am. J. Ind. Med. 28(1):99-108.

Robinson, AV. (1982) Effect of in vitro exposure to hydrogen sulfide on rabbit alveolar macrophages cultured on gas-permeable membranes. Environ. Res. 27(2):491-500.

Rogers, RE; Ferin, J. (1981) Effect of hydrogen sulfide on bacterial inactivation in the rat lung. Arch. Environ. Health 36(5):261-264.

Roth, SH; Skrajny, B; Bennington, R; et al. (1997) Neurotoxicity of hydrogen sulfide may result from inhibition of respiratory enzymes. Proc. West Pharmacol. Soc. 40(3):41-43.

Roth, SH. (2002) Personal communication to the U.S. EPA Chemical Manager.

Saillenfait, AM; Bonnet, P; de Ceaurriz, J. (1989) Effects of inhalation exposure to carbon disulfide and its combination with hydrogen sulfide on embryonal and fetal development in rats. Toxicol. Lett. 48:57-66.

Shusterman, D. (1992). Critical review: the health significance of environmental odor pollution. Arch. Environ. Health 47:76-87.

Simson, RE; Simpson, GR. (1971) Fatal hydrogen sulphide poisoning associated with industrial

waste exposure. Med. J. Aust. 1;331-334

Skrajny, B; Hannah, RS; Roth, SH. (1992) Low concentrations of hydrogen sulphide alter monoamine levels in the developing rat central nervous system. Can. J. Physiol. Pharmacol. 70(11):1515-1518.

Skrajny, B; Reiffenstein, RJ; Sainsbury, RS; et al. (1996) Effects of repeated exposures of hydrogen sulphide on rat hippocampal EEG. Toxicol. Lett. 84:43-53.

Smith, RP; Abbanat, RA. (1966) Protective effect of oxidized glutathione in acute sulfide poisoning. Toxicol. Appl. Pharmacol. 9:209-217.

Smith, RP; Gosselin, RE. (1964) The influence of methemoglobinemia on the lethality of some toxic anions. II. Sulfide. Toxicol. Appl. Pharm. 6:584-592.

Snyder, JW; Safir, EF; Summerville, GP; et al. (1995) Occupational fatality and persistent neurological sequelae after mass exposure to hydrogen sulfide. Am. J. Emerg. Med. 13: 199-203

Struve, MF; Brisbois, JN; James, RA; Marshall, MW; Dorman, DC. (2001) Neurotoxicological effects associated with short-term exposure of Sprague Dawley rats to hydrogen sulfide. Neutotoxicology 22(3):375-385.

Subramaniam, RP; Richardson, RB; Morgan, KT; Guilmette, RA and Kimbell, JS (1998) Computational fluid dynamics simulations of inspiratory airflow in the human nose and nasopharynx. Inhal. Toxicol. 10: 91-120.

Tvedt, B; Edland, A; Skyberg, K; et al. (1991a) Delayed neuropsychiatric sequelae after acute hydrogen sulfide poisoning: affection of motor function, memory, vision and hearing. Acta. Neurol. Scand. 84:348-351.

Tvedt, B; Skyberg, K; Aaserud, O; et al. (1991b) Brain damage caused by hydrogen sulfide: a follow-up study of six patients. Am. J. Ind. Med. 20:91-101.

U.S. EPA (U.S. Environmental Protection Agency). (1986a) Guidelines for the health risk assessment of chemical mixtures. Federal Register 51 (185):34014-34025.

U.S. EPA. (1986b) Guidelines for mutagenicity risk assessment. Federal Register 51(185):34006-34012.

U.S. EPA. (1988) Recommendations for and documentation of biological values for use in risk assessment. Prepared by Environmental Criteria and Assessment Office, Office of Health and Environmental Assessment, Cincinnati, OH. ERPA 600/6-87/008. Available from: National Technical Information Service, PB88-179874/AS.

U.S. EPA. (1991) Guidelines for developmental toxicity risk assessment. Federal Register 56(234):63798-63826.

U.S. EPA. (1994a) Interim policy for particle size and limit concentration issues in inhalation toxicity: notice of availability. Federal Register 59(206):53799.

U.S. EPA. (1994b) Methods for derivation of inhalation reference concentrations and application of inhalation dosimetry. Prepared by Environmental Criteria and Assessment Office, Office of Health and Environmental Assessment, Research Triangle Park, NC. EPA/600/8-90/066F.

U.S. EPA. (1994c) Peer review and peer involvement at the U.S. Environmental Protection Agency. Signed by the U.S. EPA Administrator, Carol M. Browner, dated June 7, 1994.

U.S. EPA. (1995) Use of the benchmark dose approach in health risk assessment. EPA/630/R-94/007.

U.S. EPA. (1996) Guidelines for reproductive toxicity risk assessment. Federal Register 61(212):56274-56322. http://www.epa.gov/ORD/WebPubs/repro/

U.S. EPA. (1998) Guidelines for neurotoxicity risk assessment. Federal Register 63(93):26926-26954. http://www.epa.gov/nceawww1/neurotox.htm

U.S. EPA. (1999) Revised draft guidelines for carcinogen risk assessment. NCEA-F-0644, July, 1999, Washington, DC.

U.S. EPA. (2000a) Science policy council handbook: peer review. Prepared by the Office of Science Policy, Office of Research and Development, Washington, DC. EPA 100-B-00-001.

U.S. EPA. (2000b) Science police council handbook: risk characterization. Prepared by the Office of Science Policy, Office of Research and Development, Washington, DC. EPA 100-B-00/002

Vanhoorne, M; de Rouck, A; de Bacquer, D. (1995) Epidemiological study of eye irritation by hydrogen sulphide and/or carbon disulphide exposure in viscose rayon workers. Ann. Occup. Hyg. 39(3):307-315.

Warenycia, MW; Goodwin, LR; Benishin, CG; et al. (1989a) Acute hydrogen sulfide poisoning demonstration on selective uptake of sulfide by the brainstem by measurement of brain sulfide levels. Biochem. Pharmacol. 38(6):973-982.

Warenycia, MW; Smith, KA; Blashko, CS; et al. (1989b) Monoamine oxidase inhibition as a sequel of hydrogen sulfide intoxication increases in brain catecholamine and 5 hydroxytryptamine levels. Arch. Toxicol. 63(2):131-136.

Wasch, HH; Estrin WJ; Yip, P; et al. (1989) Prolongation of the P-300 latency associated with hydrogen sulfide exposure. Arch. Neurol. 46:902-904.

Watt, MM; Watt, SJ; Seaton, A. (1997) Episode of toxic gas exposure in sewer workers. Occup. Environ. Med. 54: 277-280.

Weedon, FR; Hartzell, A; Setterstrom, C. (1940) Toxicity of ammonia, chlorine, hydrogen cyanide, hydrogen sulphide and sulphur dioxide gases V. Animals. Contributions from Boyce-Thompson Institute 11:365-385.

Weisiger, RA; Pinkus, LM; Jakoby, WB. (1980) Thiol S-methyltransferase: suggested role in detoxication of intestinal hydrogen sulfide. Biochem. Pharmacol. 29:2885-2887.

Wetterau, H; Oerkert, W; Knape, WG. (1964) Tests for the application of dried green fodder with higher H_2S content (experiments with poultry and fattened pigs). Fetterung 5: 383-393.

APPENDIX A: EXTERNAL PEER REVIEW—SUMMARY OF COMMENTS AND DISPOSITION

The support document and IRIS summary for hydrogen sulfide have undergone both internal peer review performed by scientists within EPA and a more formal external peer review performed by scientists in accordance with EPA guidance on peer review (U.S. EPA, 1998, 2000a). Comments made by the internal reviewers were addressed prior to submitting the documents for external peer review and are not part of this appendix. Public comments were received and considered. The three external peer reviewers were tasked with providing written answers to general questions on the overall assessment and on chemical-specific questions in areas of scientific controversy or uncertainty. The reviewers made a number of editorial suggestions to clarify specific portions of the text. These changes were incorporated in the document as appropriate and are not discussed further. A summary of significant comments made by the external reviewers and EPA's response to these comments follows:

(1) General Questions for Peer Reviewers

A. General Question: Are you aware of any other data/studies that are relevant (i.e., useful for the hazard identification or dose-response assessment) for the assessment of the adverse health effects, both cancer and noncancer, of this chemical?

Comment: All three reviewers considered that the data base appears to reference/summarize the relevant studies although each reviewer did suggest some additional and recently-published studies for incorporation.

Response: The studies identified by the reviewers were reviewed and information incorporated where appropriate.

B. General Question: For the RfD and RfC, has the most appropriate critical effect been chosen (i.e., that adverse effect appearing first in a dose-response continuum)? Points relevant to this determination include whether or not the choice follows from the dose-response assessment, whether the effect is considered adverse, and if the effect and the species in which it is observed in a valid animal model for humans.

Comment: One reviewer agreed that the principal study (Brenneman et al., 2000) was adequate for the determination of the RfC. However, it was suggested that, depending on the adequacy of the study design, the Hannah and Roth (1991) study could be considered as the principal study with morphological/structural changes in Purkinje cells as the critical effect. A second reviewer

declined to answer this and other questions related to the animal studies because of a lack of expertise in this area. The third reviewer did not explicitly answer this charge, but indicated that the "brain must be given first prioritythe nasal epithelial story in ratsis not relevant." This reviewer did not offer a suggestion for addressing CNS effects in humans from the data available.

Response: The Hannah and Roth (1991) perinatal inhalation study was reconsidered for use as a principal study. This reconsideration proceeded along several fronts including a focused review by Agency scientists within the National Health and Environmental Effects Research Laboratory (NHEERL).

A primary front concerns the adverse nature of the observed effect reported by Hannah and Roth, alterations in dendritic arborization of cerebellar Purkinje neurons in developing animals. The U.S. EPA neurotoxicology guidelines provide for such observations and identify them not as clearly adverse effects but rather as a "possible indicators of a neurotoxic effect." Cerebellar-related dysfunction as a possible consequence of such a possible indicator requires further elucidation. It is also acknowledged that in the case of developmental scenarios such as those used by Hannah and Roth that predicting functional impairments from specific structural alterations is difficult due to the selective nature of the observed alterations and the dynamic self-organizing response of the developing brain to injury.

Another aspect of the reconsideration involves the dose relationship of the candidate critical effects. It is clear that from the dose-response analysis accomplished in this assessment (Sections 4.3.3., 4.4.1, and 4.5.2), both nasal tract lesions and neurologic effects occur in the same concentration range at relatively low levels of H_2S exposure, and both are clearly of relevance to humans. That the levels of H_2S associated with these effects appear to be similar for either endpoint (e.g., the possible indicator of neurotoxic effects reported by Hannah and Roth to occur at 28 mg H_2S/m^3 and the no-effect level of nasal tract lesions reported by Brenneman et al. at 14 mg H_2S/m^3) is some indication that consideration for the one will also address the other.

Based on the relative clarity of the mechanistic basis, the dose-response information at lower levels of H_2S exposure, and adverse nature of the endpoint for nasal tract lesions as compared to the neurologic endpoints, as well as the recent information suggesting olfactory tissue to be a sensitive target tissue in rats (and therefore most likely in humans), nasal tract lesions are chosen as the critical effect on which to base the noncancer dose-response assessments.

C. General Question: Have the noncancer assessments been based on the most appropriate studies? These studies should present the critical effect/cancer (tumors or appropriate precursor) in the clearest dose-response relationship. If not, what other study (or studies) should be chosen and why?

Comment: One reviewer felt that although the principal and supporting studies were appropriate for derivation of RfC a more thorough discussion of the neurotoxicity toxicity studies in both the Toxicological Review and IRIS Summary should be presented. Studies requiring further discussion were identified. It was also suggested that NOAELs and LOAELs be reported for all animal studies. The second reviewer declined to answer this question. The third reviewer felt that the rat nasal studies should be removed from consideration because they are not relevant.

Response: The discussion of the neurotoxicity studies in both the Toxicological Review and IRIS summary were strengthened with additional information and discussion of the significance of hydrogen sulfide-induced CNS effects in humans was expanded in section 6 of the Toxicological Review. Additions of text analyzing the comparative dose-response between nasal tract effects and sensitive (potential) neurologic endpoints, in particular the anatomic alterations noted in cerebellar Purkinje cells by Hannah and Roth (1991), were added to Sections 4.3.3, 4.4.1, and 4.5.2. These additions provide clarification that neurologic changes and nasal tract effects (olfactory epithelium) occur at about the same externally applied concentrations of H₂S, but that the nasal tract lesions are more clearly regarded as adverse than are the most sensitive neurologic alterations noted (i.e., the alterations in Purkinje cell arborization). All animal studies were reviewed to see if it would be instructive to identify NOAELs and LOAELs for studies other than the principal and supporting studies. The nasal lesions caused by H₂S are relevant to humans for a variety of reasons, to include (1) a sentinel response alerting individuals to the presence of H_2S_1 (2) damage caused by H_2S to the olfactory nerves causing anosmia, and (3) protection against nasal lesions serves to prevent effects elsewhere, such as the brain. Along with adverse histopathological lesions localized specifically within olfactory tissue of rats exposed to airborne H₂S, Moulin et al. (2002) reported modeled simulations of H₂S flux to this tissue that was highly correlated with the lesions. These simulations imply that human olfactory tissues could also be exposed to inhaled H₂S, and, as no compelling reason exists to suspect that human olfactory tissue would react any differently to the irritant properties of H₂S, also imply that similar effects would be observed in humans.

D. General Question: Studies included in the RfD and RfC under the heading

"Supporting/Additional studies" are meant to lend scientific justification for the designation of critical effect by including any relevant pathogenesis in humans, any applicable mechanistic information, any evidence corroborative of the critical effect, or to establish the

comprehensiveness of the data base with respect to various endpoints (such as reproductive/developmental toxicity studies). Should other studies be included under the "Supporting/Additional" category? Should some studies be removed?

Comment: The only reviewer responding to this question suggested that a more comprehensive treatment of the neurotoxicity studies in humans and animals be included in the "Supporting or Additional" category.

Response: An expanded discussion of this area was included both in the Toxicological Review and IRIS summary.

E. General Question: For the noncancer assessments, are there other data that should be considered in developing the uncertainty factors or the modifying factor? Do you consider that the data support the use of different (default) values than those proposed?

Comment: Only one reviewer responded to this question. It was suggested that a discussion of species differences in breathing style and nasal anatomy be included as there is evidence to suggest that rats are more sensitive than humans as a result of these differences. The last sentence in the last paragraph on page 5 of the summary, "For these reasons, an additional uncertainty factor of 3 was considered for incorporation ...," should be deleted. If there is concern that an RfC based on olfactory lesions is not sufficiently protective of possible neurological effects, this point should be discussed in Section I.B.3. Also, is a factor of 10 needed for sensitive populations and for subchronic exposure?

Response: The reason and purpose for the dosimetric adjustment procedures applied to the effect levels in rats (i.e., to accommodate the differences in breathing style and anatomy) were made clearer to the reader in the Toxicological Review and the IRIS Summary documents.

Based on the response to the preceding questions, the concern that an RfC based on nasal tract lesions may not be able to totally address all possibilities regarding neurological effects has been made a primary consideration in the text of the Toxicological Review and the IRIS Summary documents and is reflected in text additions to Sections 4.3.3, 4.4.1, and 4.5.2.

Although the mechanism of action eliciting the nasal tract lesions (irritation) would be expected to vary directly with the delivery of the material to the nasal tract tissues, and that interindividual variability to such a direct mechanism would be minimal, the information needed to consider this matter (and a concomitant reduction of uncertainty) is not available to sufficiently compel such a reduction at this time. It is also questionable as to whether such a

direct mechanism would promote effects at lower concentration, even for much longer (i.e., lifetime) exposures, the precept on which the subchronic to chronic uncertainty factor is evoked. However, the same response has to be made for the interindividual UF, i.e., the information needed to consider this matter (and a concomitant reduction of uncertainty) is not available to sufficiently compel such a reduction at this time.

The results of Moulin et al. (2002) and Brenneman et al. (2000) indicate that olfactory tissues in rats are sensitive to the effects of H_2S , more so than respiratory epithelium. Modeling simulations (computational fluid dynamics) in rats of specific air flows to these tissues indicate olfactory tissues to be specifically sensitive to the effects of H_2S , despite higher regional flow to the respiratory epithelium (Moulin et al., 2002). As discussed above, these simulations imply that human olfactory tissues could also be exposed to inhaled H_2S , and, as no compelling reason exists to suspect that human olfactory tissue would react any differently to the irritant properties of H_2S (the suspected mechanism of damage), it is reasonable to imply that similar effects would be observed in humans.

F. General Question: Do the confidence statements and weight-of-evidence statements present a clear rationale and accurately reflect the utility of the studies chosen, the relevancy of effects to humans, and the comprehensiveness of the data base? Do these statements make sufficiently apparent all of the underlying assumptions and limitations of these assessments? If not, what needs to be added?

Comment: Only one reviewer responded to this question. It was indicated that it was unclear why there was low confidence in the RfC when the confidence in the principal study and data base was medium.

Response: The confidence statement was corrected such that the overall confidence in the RfC reflected that of the data base, i.e., medium to high.

2) Chemical-Specific Questions for Peer Reviewers

G. Specific Question: Should the reproductive toxicity of hydrogen sulfide, as seen in the Dorman et al. (2000) study be given more weight in the determination of the principal study for RfC determination?

Comment: Only one reviewer responded this question, indicating that this question was unanswerable as the effect seen on testicular tubules in rats was not clearly adverse and not confirmed in another species.

Response: A statement to this effect was added to the text of the Toxicological Review and the IRIS Summary in the study description of the Dorman et al. (2000) study.

H. Specific Question: Can you reconcile the controversy over whether hydrogen sulfide affects the brain directly with H_2S -induced brain damage due to cytochrome oxidase enzyme inhibition causing morbidity and mortality (Warenycia et al., 1989a,b) versus the peripheral (lung-related) effect of H_2S that ultimately results in hypoxia due to apnea (Baldelli et al., 1993; Almeida and Guidotti, 1999)?

Comment: One reviewer indicated that the cytochrome oxidase hypothesis and the overall weight of scientific evidence indicate that the instantaneous collapse seen upon exposure to 750-1,000 ppm is a result of H_2S directly on the brain. There must be an exposure to H_2S lasting for a minimum of about 5 minutes during a period of unconsciousness before hypoxic brain damage occurs. There is no reliable evidence that intermittent exposure to levels of H_2S below 1 or 2 ppm constitutes a danger to physical health.

Response: This effect progression was more clearly expounded upon in the text of the Toxicological Review. The increased emphasis placed in the assessment on the role of cytochrome oxidase in the progression of toxicity, and the advancement of the concept that such an alteration (i.e., statistically significant inhibition of cytochrome oxidase activity) at low levels of H_2S could be considered a sensitive indicator and harbinger of H_2S toxicity should now be apparent to the reader of this assessment.